

Removal of residual phenols from coke wastewater by adsorption

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Abstract

After biological treatment, coke wastewater contains small amounts of phenolic compounds resistant to such treatment. The removal of phenols and COD from coke wastewater subjected to biological treatment was studied. The adsorbents used were granular activated carbon and the resins XAD-2, AP-246 and OC-1074. Equilibrium, kinetics and column assays were carried out, fitting the equilibrium data to Langmuir and Freundlich models and the kinetic data to the Lagergren equation. The best results were obtained with GAC, which presented higher adsorption capacities. In the equilibrium assays, the adsorption capacities (Q) found were 1.48 mg g^{-1} for GAC versus 0.07 and 0.04 mg g^{-1} for resins AP-246 and OC-1074, respectively. In the kinetic assays, the values of the Lagergren adsorption parameter, q_e , were 1.69 , 0.15 and 0.14 mg g^{-1} for GAC, AP-246 and OC-1074, respectively. In the column assays, the dynamic capacity of GAC for up to 480 bed volumes was 1.82 mg mL^{-1} . No saturation was obtained for this volume due to the asymptotic shape of the breakthrough curve, whereas for the same percolated volume, the resins AP-246 and OC-1074 were saturated. These two resins presented similar saturation capacities of around 1.1 mg mL^{-1} .
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1. Introduction

Coke and coke by-products are produced via the pyrolysis of suitable grades of coal. The process also includes the processing of coke oven gas to remove tar, ammonia (recovered as ammonium sulphate), phenol, naphthalene, light oil, and sulphur before being used as fuel for heating the coke ovens. Coke wastewater originates from the ammonia scrubbing process.

Coke wastewater is a complex industrial wastewater present in most steelworks that contains considerable amounts of toxic compounds such as CN^- , SCN^- and phenols, as well as high concentrations of ammonium nitrogen and chlorides, though low concentrations of heavy metals and phosphorus [1,2]. It also contains small amounts of polyaromatic hydrocarbons and heterocyclic nitrogenous compounds [3–5]. Table 1 shows some typical values of the main pollutants in coke wastewater [6].

Abbreviations: COD, chemical oxygen demand; BV, bed volume; GAC, granular activated carbon; CN^- , soluble cyanide; SCN^- , thiocyanate; NO_2^- , nitrite; NO_3^- , nitrate; Cl^- , chloride

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Individual concentrations of each component vary according to the type of coal used and modifications of each specific process [7].

Traditionally, expensive processes have been employed for coke wastewater treatment, such as the addition of caustic soda and steam stripping. More recently, biological treatments have been applied to reduce the pollutant load. Several treatment plants have already been built in the German district of Rhine Westphalia. Some processes, such as NITE/DENITE, ENSR Corp., Houston, TX patented, are based on biological treatment of coke wastewater without any prior pre-treatment, since no ammonium stripping is carried out. The best known biological treatment plants in Europe are Zentralkokerei Saar and Kaiserstuhl in Germany, Serémage in France and Sidmar in Belgium. In Spain, the first biological plant was built at the end of 2003 at the Arcelor group facilities in Avilés (Asturias). This plant consists in a modification of a conventional activated sludge process using pure oxygen instead of air with an anoxic area for denitrification.

After this treatment, the concentration of phenols in the wastewater exceeded the limits established in both state regulations regarding direct discharge into watercourses (Royal Decree 606/2003) as well as regional regulations (Law 5/2002)

Nomenclature

C	substrate concentration
C_e	the equilibrium concentration of the species in the solution (mg L^{-1})
C_0	substrate initial concentration
k	constant of the Langmuir equation related to the enthalpy of the process (mg L^{-1}) ⁻¹
k'	the equilibrium rate constant (min^{-1})
K_F	Freundlich surface adsorption equilibrium constant (mg L^{-1}) ⁻¹
n	heterogeneity factor
q_e	amount sorbed per specified amount of sorbent (mg g^{-1})
q_t	amount of substrate sorbed at time t (mg g^{-1})
Q	adsorption capacity to form the single layer (mg g^{-1})

Table 1
Typical pollutant concentrations in coke wastewater

Parameter	Coke wastewater		
	Australia	Germany	Spain
BOD ₅ (mg L^{-1})	610	1600–2600	1150
COD (mg L^{-1})	2200	4000–6500	3030
TSS (mg L^{-1})	50	2–10	31
TKN (mg L^{-1})	330	300–500	1245
NH ₄ ⁺ -N (mg L^{-1})	272	50–150	1010
P (mg L^{-1})	<1	<1	<1
Phenol (mg L^{-1})	333	400–1.200	485
SCN ⁻ (mg L^{-1})	184	200–500	215
CN ⁻ (mg L^{-1})	93	4–15	50

regarding discharge into public sewage systems, which means that a polishing treatment is required.

Phenols can be highly toxic even at low concentrations and their concentrations in aquatic media are strictly limited by regulations. In recent years, several physico-chemical treatments have been proposed for obtaining efficient phenol removal, such as adsorption, air stripping, steam stripping, wet oxidation, heterogeneous photocatalysis and biological treatments [8–14].

In this paper, the possibility of using adsorbents for the removal of residual concentrations of phenols and COD in coke wastewater after biological treatment is studied.

2. Materials and methods

2.1. Coke wastewater characteristics

The coke wastewater used in this study came from the Arcelor plant in Aviles (Spain), which produces 1512 m³ d⁻¹ of wastewater (0.38 m³ t⁻¹ coke). Its physico-chemical characteristics are shown in Table 2. The main pollutants present are ammonium nitrogen (650–700 mg L⁻¹), phenols (180–250 mg L⁻¹), thiocyanates (200–300 mg L⁻¹) and chemical oxygen demand

Table 2

Average pollutant concentration in coke wastewater before biological treatment

Parameter	Average concentration
pH	8.1
Conductivity (mS cm^{-1})	7.1
COD ($\text{mg O}_2 \text{L}^{-1}$)	1100
BOD ₅ ($\text{mg O}_2 \text{L}^{-1}$)	579
TSS (mg L^{-1})	32.0
VSS (mg L^{-1})	23.0
Phenols (mg L^{-1})	207
Oil and grease (mg L^{-1})	5.3
Alkalinity ($\text{g CaCO}_3 \text{L}^{-1}$)	0.25
NH ₄ ⁺ -N (mg L^{-1})	688
TKN (mg L^{-1})	757
NO ₃ ⁻ (mg L^{-1})	76.0
SCN ⁻ (mg L^{-1})	267
CN ⁻ (mg L^{-1})	32
PO ₄ ³⁻ (mg L^{-1})	0.5
SO ₄ ²⁻ (mg L^{-1})	102
S ²⁻ (mg L^{-1})	0.1
Cl ⁻ (mg L^{-1})	1290
Fe (mg L^{-1})	4.4
Mg (mg L^{-1})	6.1
Ca (mg L^{-1})	11.0

(1000–1200 mg L⁻¹), of which around 70% is made up of biodegradable matter. Another toxic pollutant present is cyanide (20–45 mg L⁻¹). The wastewater was treated in a laboratory-scale activated sludge plant with the following characteristics (Fig. 1): an anoxic reactor for the denitrification process, an aerobic reactor for the biodegradation of phenols and thiocyanate followed by a settling tank to separate the biomass and a second aerobic reactor for the nitrification process, likewise followed by a settling tank.

The average composition of the effluent from this plant is shown in Table 3. The phenol concentration ranged from 8 to 2.5 mg L⁻¹, COD from 500 to 400 mg L⁻¹, NH₄⁺-N from 30 to 40 mg L⁻¹ and very low concentrations of SCN⁻ and CN⁻. Adsorption experiments using several adsorbents were carried out in order to decrease the phenol content to values ≤ 2 mg L⁻¹, thus enabling the wastewater to be discharged into the sewage system.

2.2. Preliminary adsorption assays

Initially, several assays were conducted to select the best adsorbents. The adsorbents used were: granular activated carbon

Table 3

Average pollutant concentration in coke wastewater after biological treatment

Parameter	Average concentration
pH	8.6
COD (mg L^{-1})	430
SCN ⁻ (mg L^{-1})	1.2
NH ₄ ⁺ -N (mg L^{-1})	34
TSS (mg L^{-1})	60
Phenols (mg L^{-1})	5.3
CN ⁻ (mg L^{-1})	0.2

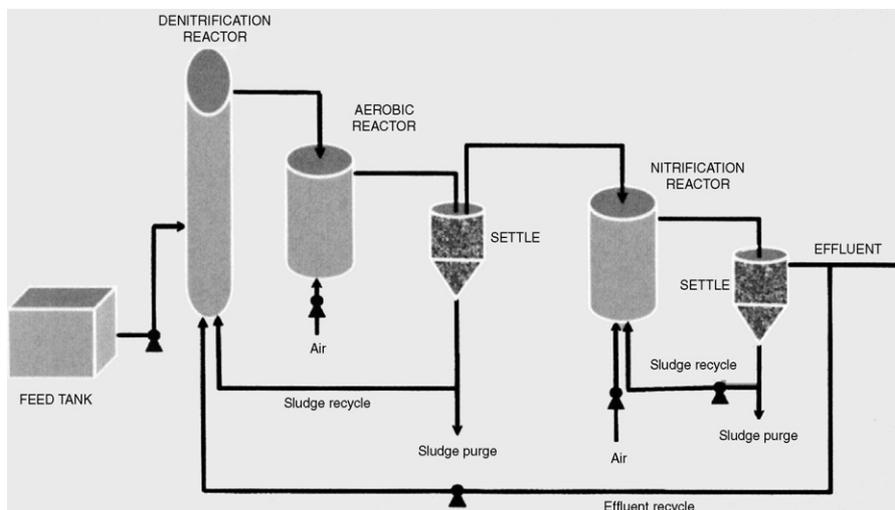


Fig. 1. Activated sludge laboratory-scale plant.

with different particle sizes (0.8, 1.0, 1.4 and 2.5 mm diameter), and the resins AP-246, XAD-2 and OC-1074, the characteristics of which are shown in Table 4.

2.3. Equilibrium and kinetic assays

The kinetic assays were carried out in 250 mL Erlenmeyer flasks sealed with Parafilm, to prevent the loss of phenol by volatilisation. In the flasks, 100 mL of phenol solution of known concentration (5, 10 and 15 mg L⁻¹) and initial pH 8.4 were put in contact with different amounts of adsorbents (0.5, 1, 2 and 4 g). The phenol solutions were prepared with original wastewater coming from the pilot-scale wastewater treatment plant at Arcelor facilities. The flasks were shaken using a VIBRATIC mixer for the selected time (5, 10, 15, 20, 30, 60, 90, 120, 150 and 180 min). The assays were carried out at room temperature.

For the study of adsorption equilibrium at 20 °C, 1, 2 and 4 g of adsorbents were put in contact with 100 mL of wastewater. The initial concentration of phenol ranged from 5 to 15 mg L⁻¹ at a pH of 8.4. The flask were shaken using a VIBRATIC mixer

for 24 h. At the end of selected period time, the content of each flask was filtered through a 0.45 µm membrane filter and phenol concentration in the filtrate was determined.

2.4. Column assays

The column adsorption assays were conducted downflow in a glass column (2 cm diameter and 13 cm height) at a rate of 6.4 BV h⁻¹. The biologically treated coke wastewater had initial concentrations of phenols and COD of 5.3 and 430 mg L⁻¹, respectively. After the adsorption assays, regeneration was carried out in counter current mode at a linear velocity of 14.1 BV h⁻¹ employing 6% NaOH for the regeneration of GAC and 20% NaCl and 2% NaOH for the regeneration of the two resins.

2.5. Chemical analyses

The parameters analysed in the coke wastewater were: pH, conductivity, chemical oxygen demand (COD), biochemical oxygen demand (BOD), suspended solids (SS), volatile

Table 4
Characteristics of the adsorbent resins employed

	AP-246 WS	VPOC-1074	XAD-2
Ionic form, as shipped	Cl ⁻ /OH ⁻	Cl ⁻	Cl ⁻ /OH ⁻
Grain shape	Balls	Bead-shaped	Balls
Appearance	Light-beige, opaque	White, opaque	White, opaque
Matrix	Polyacrylamide	Polyacrylamide	Polystyrene-DVB
Functional group	Quaternary amine	Quaternary amine	
Bead size	0.3–1.25 mm	0.4–1.6 mm	0.3–1.25 mm
Effective size	0.48 mm	0.55 mm	
Bulk density	700–800 g L ⁻¹	680 g L ⁻¹	640–700 g L ⁻¹
Density	1.08	1.08	1.06 g mL ⁻¹
Water retention	60–65%	65–72%	
Stability at pH range	1–14	0–14	1–14

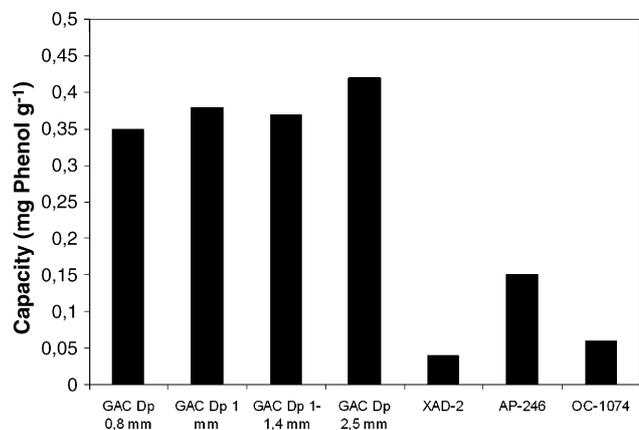


Fig. 2. Results of the batch preliminary adsorption assays.

suspended solids (VSS), phenols, oil and grease, alkalinity, Kjeldahl nitrogen, ammonium nitrogen, nitrate, thiocyanate, phosphate, cyanide, sulphate, sulphide, chloride and metals (Fe, Mg and Ca).

Phenols and COD were determined for the adsorption study. All the analyses were performed according to Standard Methods [15].

3. Results and discussion

3.1. Preliminary assays

The results can be seen in Fig. 2. In general, GAC shows higher adsorption capacities than the resins. GAC capacities range from 0.35 mg g⁻¹ for 0.8 mm diameter GAC to 0.45 mg g⁻¹ for 2.5 mm diameter GAC. Among the resins, the best adsorption capacity was obtained with the AP-246 resin (0.15 mg g⁻¹), followed by the OC-1074 resin and the XAD-2 resin.

3.2. Equilibrium and kinetic assays

In line with these results, equilibrium and kinetic assays were performed with GAC 2.5 mm diameter and the AP-246 and OC-1074 resins. The equilibrium assays were fitted to Langmuir and Freundlich models. The Langmuir isotherm fits the following

equation:

$$q_e = \frac{QkC_e}{1 + kC_e}$$

where q_e is the amount sorbed per specified amount of sorbent (mg g⁻¹), Q the sorption capacity to form the single layer (mg g⁻¹), k the constant of the Langmuir equation related to the enthalpy of the process (mg L⁻¹)⁻¹, and C_e is the concentration of the species in solution (mg L⁻¹).

This isotherm is applicable under the following hypothesis: (i) the solid has a uniform surface, (ii) absence of interactions between the solid molecules, and (iii) the sorption process takes place in a single layer.

The Freundlich isotherm fits the following equation:

$$q_e = K_F C_e^{1/n}$$

where q_e is the amount sorbed per specified amount of adsorbent (mg g⁻¹), K_F the constant related to the sorption capacity (mg L⁻¹)⁻¹, C_e the equilibrium concentration in solution (mg g⁻¹), and n is the empirical parameter related to the intensity of sorption, which varies with the heterogeneity of the material. When $1/n$ values are in the range $0.1 < 1/n < 1$, the sorption process is favourable.

This model is valid for heterogeneous surfaces and predicts an increase in the concentration of the ionic species sorbed onto the surface of the solid with increasing concentrations of said species in the liquid phase.

The calculated Langmuir and Freundlich parameters are presented in Table 5. When applying the Langmuir model, the maximum phenol adsorption capacity in a single layer was obtained for GAC (1.48 mg g⁻¹), the adsorption capacities of the resins being much lower (0.07 and 0.04 mg g⁻¹ for AP-246 and OC-1074, respectively). When fitted to the Freundlich equation, the value of K_F was higher for GAC, 0.97 (mg L⁻¹)⁻¹, and much lower for the resins. The $1/n$ value was slightly unfavourable for GAC, with a value of 1.26, whereas the values clearly indicate an unfavourable equilibrium for the resins.

The kinetic results were fitted to the Lagergren equation, based on solid capacity. The Lagergren equation is one of the most widely used sorption rate equations for the sorption of a solute from a liquid solution. It may be expressed as:

$$\frac{dq_t}{dt} = k'(q_e - q_t)$$

Table 5
Langmuir and Freundlich parameters for phenol and COD adsorption onto GAC (2.5 mm) and resins AP-246 and OC-1074

	Langmuir parameters			Freundlich parameters		
	Q (mg g ⁻¹)	K (mg L ⁻¹) ⁻¹	r	K_F (mg g ⁻¹)	$1/n$	r
Phenol adsorption						
Activated coal	1.481	0.065	0.98	0.97	1.261	0.97
Resin AP-246	0.071	0.584	0.98	0.112	2.830	0.98
Resin OC-1074	0.043	0.445	0.99	0.0053	6.141	0.99
COD adsorption						
Activated coal	74.626	0.0023	0.99	1.483×10^{-6}	3.258	0.99
Resin AP-246	0.423	0.0036	0.99	2.15×10^{-36}	15.102	0.98
Resin OC-1074	3.559	0.0021	0.96	2.90×10^{-8}	3.352	0.96

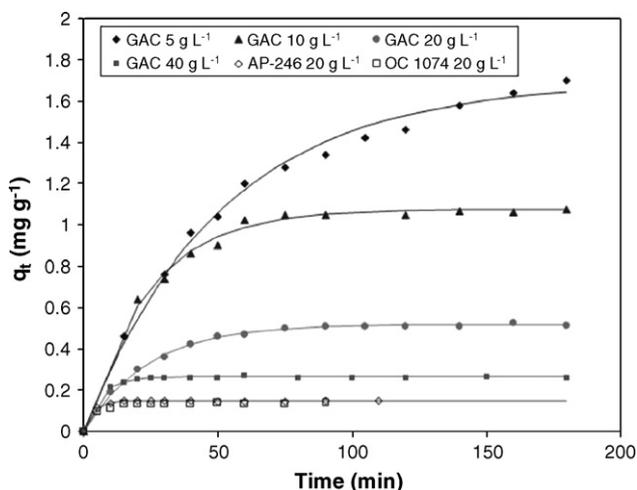


Fig. 3. Kinetics of phenol adsorption on GAC fitted to Lagergren equation.

Integrating this equation between $t=0$ and t and $q_t=0$ and q_t , we obtain:

$$\log \left(\frac{q_e}{q_e - q_t} \right) = \frac{k'}{2.303} t$$

This is the integrated rate law for a pseudo-first order reaction, where q_e is the amount of substrate sorbed at equilibrium (mg g^{-1}), q_t the amount sorbed at time t (mg g^{-1}) and k' is the equilibrium rate constant (min^{-1}) [16].

The experimental data fitted to the Lagergren equation are shown in Fig. 3 and Table 6. In all cases, the regression coefficient was higher than 0.996. The q_e values obtained for an initial adsorbent concentration of 20 g L^{-1} were 0.52 , 0.15 and 0.14 mg g^{-1} for GAC, resin AP-246 and resin OC-1074, respectively. Experiments with different adsorbent dosages, performed using GAC, clearly show that when the dosage is increased, the adsorption capacity decreases, as expected. The capacity decreased from 1.69 mg g^{-1} for 5 g GAC L^{-1} to 0.26 mg g^{-1} for 40 g GAC L^{-1} .

With respect to the influence of phenol concentration, the results obtained, shown in Fig. 4, indicate that when the phenol concentration increases, the adsorption capacity also increases for the same dosage of activated carbon (40 g L^{-1}). The q_e

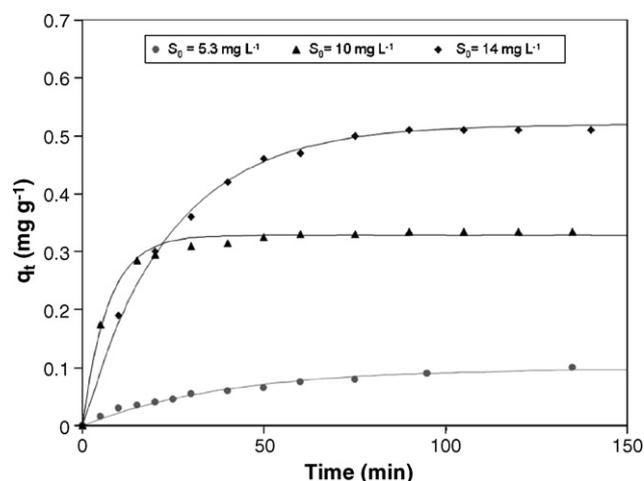


Fig. 4. Influence of phenol concentration on the adsorption capacity of GAC.

increased from 0.1 to 0.5 mg g^{-1} when the phenol concentration changed from 5.3 to 14 mg L^{-1} .

3.3. Column assays

Figs. 5–7 show the breakthrough curves obtained in the column assays using the different adsorbents. In keeping with the results obtained in the equilibrium and kinetic assays, GAC showed the best capacity for phenol adsorption. It was possible to treat 450 BV , obtaining an effluent concentration $\leq 1.6 \text{ mg L}^{-1}$ ($C/C_0=0.30$); for 480 BV , the adsorption capacity was 1.82 mg mL^{-1} (asymptotic curve).

The AP-246 and OC-1074 resins presented poorer adsorption characteristics; when treating 200 – 250 BV , the phenol concentration in the effluent was 1.9 – 2.1 mg L^{-1} ($C/C_0=0.35$ – 0.4). Saturation capacities were similar for both resins, around 1.1 mg mL^{-1} .

The adsorption of COD was much lower than the adsorption of phenol, the effluent reaching a concentration of 344 mg L^{-1} ($C/C_0=0.8$) with 100 BV for GAC and with 50 BV for the resins.

The results of the regeneration of the three adsorbents are shown in Fig. 8. Around 20 BV were needed to regenerate the

Table 6
Lagergren parameters for phenol adsorption onto GAC (2.5 mm) and resins AP-246 and OC-1074

	Kinetic equation		
	q_e	k'	r
Activated coal			
5 g L^{-1}	1.693	0.0195	0.9968
10 g L^{-1}	1.076	0.0419	0.9974
20 g L^{-1}	0.519	0.0430	0.9990
40 g L^{-1}	0.264	0.1624	0.9985
Resin AP-246			
20 g L^{-1}	0.149	0.2897	0.9975
Resin OC-1074			
20 g L^{-1}	0.137	0.2246	0.9963

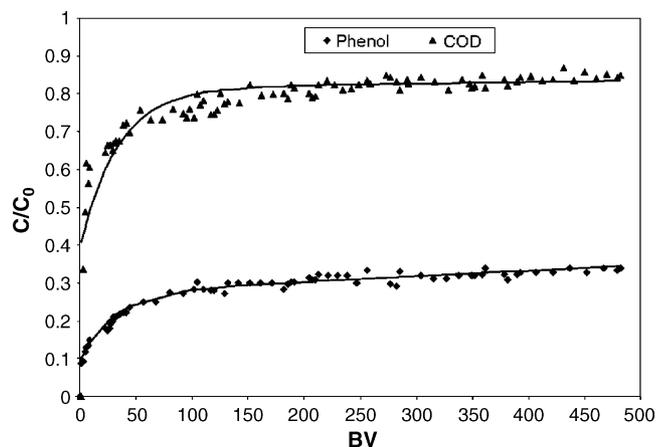


Fig. 5. Breakthrough curves obtained with GAC.

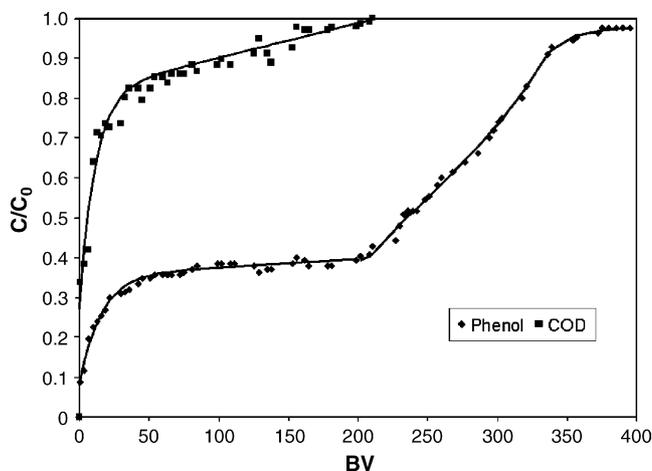


Fig. 6. Breakthrough curves obtained with resin AP-246.

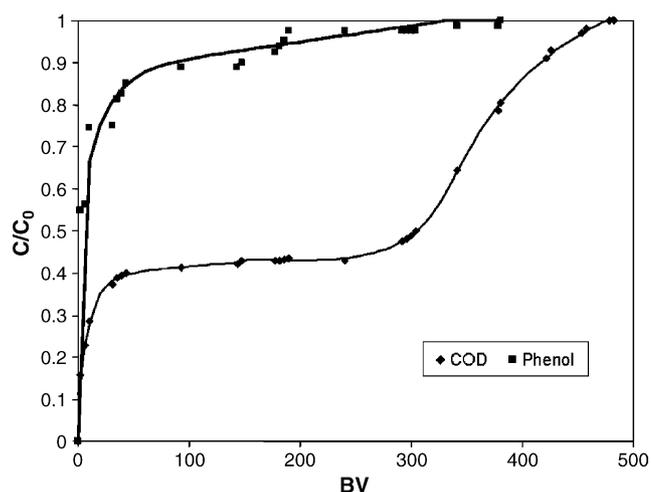


Fig. 7. Breakthrough curves obtained with resin OC-1074.

resins, although the regeneration capacity was not very good: 64.5% for OC-1074 and 38.1% for AP-246. It was not possible to regenerate the GAC with 6% NaOH (the regeneration capacity was only 3%). However, due to the lower price of this adsorbent and to the loss in properties when regeneration is carried out (reduction of the pore surface), the current ten-

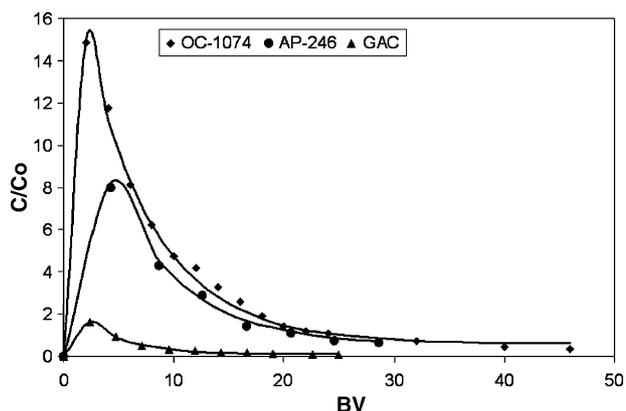


Fig. 8. Regeneration curves for the adsorbents GAC, AP-246 and OC-1074.

dustry is the disposal or incineration of this adsorbent after its exhaustion.

4. Conclusion

Residual phenols from high load industrial wastewater biological treatment plants can be efficiently removed using activated carbon. The concentration of pollutants after adsorption permits the effluent to be discharged into the sewage system for subsequent treatment at a sewage plant. The other adsorbents tested in this study (XAD-2, AP-246 and OC-1074 resins) were less adequate, especially XAD-2, due to their lower adsorption capacities.

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References

- [1] J. Wanmmer, The implementation of bulking control in design of activated sludge systems, *Water Sci. Technol.* 29 (7) (1994) 193–202.
- [2] G.M. Wong-Chong, Retrofitting LTV coke plant wastewater treatment system to comply with pretreatment discharge limits, *Iron Steel Eng.* (1994) 26–28.
- [3] D. Bamelis, Biological purification of coking plant waters, *Rev. Metall.* (1992) 132–145.
- [4] D. Jenkins, Towards a comprehensive model of activated sludge bulking and foaming, *Water Sci. Technol.* 25 (6) (1992) 215–230.
- [5] W. Keith, J. Antil, Biotreatment of coke oven effluent, *Steel Time Int.* (1991) 26–35.
- [6] I. Vázquez, Estudio y modelización de la eliminación biológica de contaminantes en aguas de coquería, Doctoral Thesis, University of Oviedo, 2005.
- [7] S. Schacht, J. Klein, Biological removal of pollutants. New process for the treatment of coking plant wastewater, *Energie* 41 (10) (1989) 32–37.
- [8] G. González, G. Herrera, Ma.T. García, M. Peña, Biodegradation of phenolic industrial wastewater in a fluidized bed reactor with immobilized cells of *Pseudomonas putida*, *Bioresour. Technol.* 80 (2001) 137–142.
- [9] G. González, G. Herrera, Ma.T. García, M. Peña, Biodegradation of phenol in a continuous process: comparative study of stirred tank and fluidised-bed bioreactors, *Bioresour. Technol.* 76 (2001) 245–251.
- [10] S.T. Christoskova, M. Stoyanova, Degradation of phenolic wastewaters over Ni-oxide, *Water Res.* 35 (8) (2001) 2073–2077.
- [11] C.S.A. Sá, R.A.R. Boaventura, Biodegradation of phenol by *Pseudomonas putida* DSM 548 in a trickling bed reactor, *Biochem. Eng. J.* 9 (2001) 211–219.
- [12] C. Aharoni, C.L. Sparks, Kinetics of soil chemical processes: a theoretical treatment, in: D.L. Sparks, D.L. Suárez (Eds.), *Rates of Soil Chemical Processes*. Soil Sci. Soc. Am. Madison, 1991, pp. 1–18.
- [13] Q. Wu, X. Hu, P.L. Yue, Kinetics study on catalytic wet air oxidation of phenols, *Chem. Eng. Sci.* 58 (2003) 923–928.
- [14] C.B. Maugans, A. Akgerman, Catalytic wet oxidation of phenol in a trickle bed reactor over a Pt/TiO₂ catalyst, *Water Res.* 37 (2003) 319–328.
- [15] APHA, AWWA, WEF, *Standard Methods for the Examination of Water and Wastewater*, 20th ed., APHA/AWWA/WEF, Washington, DC, 1998.
- [16] I. Alemzadeh, F. Vossoughi, M. Houshmande, Phenol biodegradation by rotating biological contactor, *Biochem. Eng. J.* 11 (2002) 19–23.