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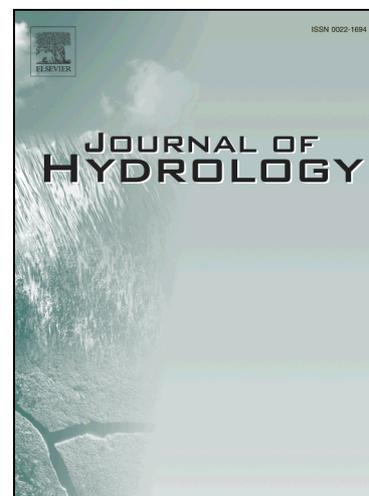
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1 Evaluating the reliability of stormwater treatment systems under various 2 future climate conditions

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13 Abstract

14 Water Sensitive Urban Design (WSUD) stormwater systems, also known as Low Impact
15 Development (LID) systems or Nature Based Solutions (NBS), are currently implemented based on
16 the underlying assumption of statistical stationarity of rainfall, which threatens to become outdated
17 under climatic uncertainty. This paper applies a new downscaling method to examine the implications
18 of climate change on future rainfall and evaluate the reliability of WSUD stormwater infrastructure in
19 pollution reduction, flow frequency mitigation and reliability as an alternative water supply. A variety
20 of future atmospheric scenarios are considered as part of this comprehensive assessment by analysing
21 an ensemble of eight different downscaled General Circulation Models (GCMs). High resolution
22 catchment-scale rainfall projections for Melbourne, Australia were generated using a scheme called
23 High-resolution Downscaling of Rainfall Using STEPS (HiDRUS) at a fine 1 km and 6-minute scale
24 for more precise analysis with uncertainty estimates. Statistical analyses show that, in general, the
25 climate models predict a drier future with fewer rainfall events and longer dry periods when
26 comparing the simulated near future (2040-2049) periods against the base-line period (1995-2004).
27 The difference simulated between historical and future rainfall projections show minimum difference
28 of WSUD performance in pollution removal and flow frequency reduction, with slightly lower
29 harvesting reliability (<3%) observed under future climate; high variabilities, however, were observed
30 across GCM simulations, indicating big uncertainties of system reliability under various conditions,
31 e.g. design wetland sizes may vary from 2.5% to 4.0% of the impervious catchment area according to

32 different future projects across GCMs. Larger WSUD systems are recommended to ensure reliable
33 performance of pollution removal, as well as harvesting reliability under simulated future conditions.
34 The significance of considering an ensemble of different GCMs as opposed to many scenarios
35 generated by a single 'best' climate model was also demonstrated for the robust estimation of
36 uncertainty in future WSUD reliability. This work highlights important considerations for the future
37 design, management and quantitative evaluation of WSUD reliability.

38 **Keywords:** General Circulation Model (GCM); climate downscaling; stormwater management;
39 pollutant removal; stormwater harvesting; Water Sensitive Urban Design (WSUD)

40 1. Introduction

41 Urbanisation and climate change are growing concerns (Semadeni-Davies *et al.*, 2008; IPCC, 2013)
42 and pose challenges to water professionals. Numerical studies have been conducted to investigate the
43 impact of future climate and rapid urbanisation on urban catchment hydrology requiring fine
44 resolution data (e.g. Andréasson *et al.*, 2004; Denault *et al.*, 2006; Olsson *et al.*, 2009; Zahmatkesh *et*
45 *al.*, 2015). Future climate scenarios with fine resolutions can be generated in many ways, via
46 artificially decreasing/increasing historical rainfall time series (e.g. Bach *et al.*, 2013; Urich *et al.*,
47 2013), or through adjusting Intensity-Frequency-Duration (IFD) curves (e.g. McIntyre *et al.*, 2007);
48 different downscaling techniques, including dynamic downscaling (e.g. Thoeun, 2015), and statistical
49 downscaling (e.g. Hewitson and Crane, 1996; Rummukainen, 1997) have also been used to generate
50 future projections from GCMs. These methods are often used to investigate the climate change on
51 urban hydrological processes (Denault *et al.*, 2006; Olsson *et al.*, 2009) and conventional urban
52 drainage systems (Prudhomme *et al.*, 2002; Semadeni-Davies *et al.*, 2008) in a regional / city scales. It
53 is frequently found that higher rainfall intensities and peak flows are expected in urban areas in the
54 future thereby impacting the performance of conventional urban drainage networks with larger flows,
55 longer event durations and more frequent pluvial floods (Ashley *et al.*, 2005; Rosenberg *et al.*, 2010;
56 Kang *et al.*, 2016; Wang *et al.*, 2017). The studies regarding the impact of climate change on local
57 catchment scales are often limited due to the lack of finer scale of space-time rainfall predictions.

58 Recently, Raut *et al.* (2018) developed a multiplicative random cascade model – HiDRUS, which is
59 able to generate very fine spatial and temporal resolution rainfall projections (in one kilometre and six
60 minutes); HiDRUS is the first kind of method that allows for GCM outputs to be downscaled to a
61 local catchment scale, taking into account the topographical features of the local landscape.

62 Sustainable solutions, *e.g.* stormwater biofilters, constructed wetlands and ponds, developed under the
63 concept of Water Sensitive Urban Design (WSUD, also known as LID or NBS) (Fletcher *et al.*,
64 2015), were designed to mimic natural hydrological processes removed through urbanisation
65 (Semadeni-Davies *et al.*, 2008), and they have also been found to effectively counter the adverse
66 effects of climate change and urbanisation (Bach *et al.*, 2013; Wang *et al.*, 2017). However, these
67 systems are traditionally designed and operated under the basis of stationary (*e.g.* using historical
68 data) to achieve required treatment targets (VSC, 1999). This, despite the fact that climate change
69 leads to more extreme rainfall (across most of Australia) and longer dry periods (in southeast and
70 southwest Australia) (Steffen *et al.*, 2017), can impact the structural integrity and function of many of
71 these natural systems; for example extreme events with either long dry periods or large volumes were
72 found to adversely impact the pollution removal performance of stormwater biofilters (Zhang *et al.*,
73 2014). Hence, with detailed historical hydrological inputs used to design WSUD and their perceived
74 long operational lifespan (*e.g.* usually goes beyond 20 years), it is questionable whether currently
75 designed WSUD systems can still deliver adequate treatment several years later and, thus, it is
76 necessary to investigate the impact that climate change can inflict.

77 Relevant research on climate change impact on the performance of WSUD systems is rather limited
78 due to the requirement of high space-time resolution of rainfall data and it is currently unclear
79 whether existing WSUD systems can provide adequate treatment under future climate. Burge *et al.*
80 (2012) conducted a case study in Melbourne, Australia to measure the likely impact of climate change
81 on WSUDs using multiple scenarios of adjusted historical rainfall time series to represent the
82 extremes of a number of projected ranges of climate change scenarios. It was found that potential
83 climate change futures will have minimal impact on the efficiency and effectiveness of WSUD
84 infrastructure, *e.g.* most stormwater treatment devices cope very well with the climate change

85 predictions (worst case scenario was a pollutant load reduction performance of only up to 6% for
86 bioretentions systems) while, in the most cases, the reduction in system reliability as an alternative
87 water supply was within 5% of the base case. Sharma *et al.* (2016) however found that climate change
88 will lead to increased outflow concentrations of total suspended solids and copper from a stormwater
89 retention pond by comparing current climate and a synthetic future climate scenario with increased
90 intensity of rainfall events and longer dry periods. Both studies used simple approaches to account for
91 future climate (*i.e.* rainfall scaling/adjustments) and the contradicting findings trigger the further need
92 for investigating the level of uncertainty of the treatment performance of WSUD systems under future
93 climate and, hence, understanding the reliability of these systems designed according to the current
94 paradigm, against variable future scenarios.

95 As decentralised WSUD systems are usually designed to treat medium to small events (*i.e.* up to 2
96 years return periods) and can be located at the smaller street and allotment scales, higher spatial (~
97 1 km) and temporal resolution (< hourly) of future rainfall time series are needed (Ochoa-Rodriguez
98 *et al.*, 2015). However, the lack of fine catchment-scale future rainfall time-series impedes the
99 advancement of our understanding of WSUD system reliability under future climate. A further
100 impediment is that there are even fewer studies that consider multiple scenarios in accounting for
101 future climate uncertainties (Jones, 2000). GCMs are diverse in their implementation of the physics of
102 atmosphere and have different couplings of atmosphere-ocean-ice components (IPCC, 2013).
103 Therefore, it is reasonable to hypothesise that investigating multiple scenarios across GCMs would be
104 far superior to exploring various scenarios within a single specific GCM (e.g. Rosenberg *et al.*, 2010);
105 this can also lead to as a more conservative all-encompassing understanding about the degree of
106 uncertainty of future system reliability for WSUD planning and robust decision making.

107 The aim of this study is to investigate the impact of future climate for three key functions of WSUD:
108 (1) pollutant treatment performance and (2) flow frequency reduction (both are investigated for two
109 common decentralised WSUD technologies – biofilter systems and constructed wetlands), as well as
110 (3) harvesting reliability (based on the design of rainwater tanks). To undertake this study, we used
111 HiDRUS (Raut *et al.*, 2018) to downscale fine spatial and temporal resolution rainfall projections (in

112 one kilometre and six minutes) from an ensemble of eight different GCMs. Statistical investigation of
113 historical and future rainfall projections for different rainfall characteristics was performed and
114 compared to rain gauge data. We identified the extent of the future rainfall variability and the extent
115 to which multiple rainfall time series from a single GCM vs. an ensemble of time series from multiple
116 GCMs should be considered. Using an ‘optimum number’ of simulations extracted from different
117 GCMs, we then propagated these rainfall projections to WSUD models to investigating their impact
118 on system reliability and the associated design variability.

119 To the authors’ knowledge, this is the first study that evaluates the future reliability of WSUD systems
120 using finer scale space-time predictions of future rainfall downscaled from GCMs against a multitude
121 of system behaviours. This study specifically focuses on Melbourne, Australia as the city has a rich
122 history with WSUD and thousands of assets already implemented (Kuller *et al.*, 2018). Nevertheless,
123 the methods are state-of-the-art and transferable to any other region in the world for which future
124 rainfall data can be obtained. It should be acknowledged that our goal is not to validate whether
125 GCMs and the HiDRUS can reproduce exact historical rainfall patterns, but to investigate variabilities
126 of WSUD reliability impacted by future climate.

127 **2. Methods**

128 **2.1 Rainfall data sets**

129 This study specifically focused on one site: Melbourne Regional Office (MRO; Latitude: -37.81 °S
130 Longitude: 144.97 °E), which has observed 6-min time series rainfall data for 1995-2004 (from the
131 Bureau of Meteorology – BoM). The observed data is regarded as the baseline for the evaluation of
132 climate change effects. Ten-year periods have been found to achieve a satisfactory compromise
133 between modelling a short period that represents current climate conditions and a longer period that
134 better illustrates climatic variability (Ashley *et al.*, 2005). In the BoM rainfall data, there were two
135 periods of missing data: 1-30 Nov 1998 and 1-28 Feb 2002; the data in the missing period were
136 recorded as zeros but replaced by the authors using data from a nearby rain gauge station (Bundoora;
137 Latitude: -37.72 °S Longitude: 145.05 °E; about 18 km north-east from MRO).

138 In this study, eight different GCMs that span the entire variability of the Australian region were
 139 selected (Table 1); a similar set of GCMs was also used in other studies within this region (Raut *et al.*,
 140 2016). Selecting only eight GCMs was also a compromise due to (1) large disk space requirement
 141 for storing the downscaled rainfall data-sets that have very fine spatial and temporal resolutions (e.g.
 142 > 100 TB), and (2) long computation time for WSUD modelling using the long-term downscaled
 143 rainfall series. This study used the RCP8.5 scenario from CMIP5 which represents the CO₂ emission
 144 scenario that is highly energy-intensive as a result of high population growth and a lower rate of
 145 technology development (van Vuuren *et al.*, 2011).

146 **Table 1** GCMs used in this study: temporal resolution of all GCMs is daily.

GCM	Modeling Group, Country	Horizontal Resolution (Latitude × Longitude*)
ACCESS1-0	CSIRO, Australia	1.25° × 1.875°
ACCESS1-3	CSIRO, Australia	1.25° × 1.875°
BCC-CSM1	Beijing Climate Centre, China	1.9° × 1.9°
CMCC-CMS	Centro Euro-Mediterraneo, Italy	2° × 2°
CNRM-CM5	Meteo-France, France	1.9° × 1.9°
GFDL-CM3	Geophysical Fluid, Dynamics Lab, USA	2.0° × 2.5°
MIRO-C5	Centre for Climate System Research, Japan	1.4° × 1.4°
MRI-CGCM3	Meteorological Research Institute, Japan	2.8° × 2.8°

147 The HiDRUS model developed by Raut *et al.* (2018), was used to downscale the eight GCMs. The
 148 model uses multiplicative random cascades (which is often used in the urban hydrological context, e.g.
 149 Licznar *et al.* (2015) and Müller and Haberlandt (2018)) from the Short-Term Ensemble Prediction
 150 Systems (STEPS, Seed *et al.*, 1999) to generate high-resolution rainfall fields. The rainfall structures
 151 follow Lagrangian temporal evolution in an AR2 framework. For details on the model, please refer to
 152 Seed *et al.* (1999), Raut *et al.* (2012) and Raut *et al.* (2018). Briefly, the model is capable of
 153 downscaling rainfall from time and space resolutions of several hours and a hundreds of kilometres to
 154 scales of minutes and kilometres (which is required for catchment-scale modelling). The model can
 155 successfully reproduce the frequency distribution of 6-minute rainfall intensities, storm durations,
 156 inter-arrival times and autocorrelation function against radar data at 12 locations in the Greater
 157 Melbourne area (Raut *et al.*, 2018). The spatial variation in rainfall accumulation is realised by using
 158 multiplicative bias factors computed from the observed data (Raut *et al.*, 2018). The model was run
 159 with historical parameters estimated from the Melbourne radar data during the observation period
 160 2008-2015. It should be noted that the period/data used to estimate historical parameters (2008-2015;

161 radar data) is different from the historical period of this current study (1995-2004; rain gauge data);
162 further studies by Raut *et al.* (Submitted manuscript) on testing HiDRUS has validated that the
163 parameter transferability was acceptable across different periods and datasets.

164 Using HiDRUS, 100 continuous simulations of rainfall at 1 km spatial and 6 minutes temporal
165 resolutions were generated from each GCM over the periods of 1995-2004 (historical projections) and
166 2040-2049 (“near-future” projections – for simplification “future” is used throughout the paper
167 instead of “near-future”) at MRO. The 100 different simulations provide estimation of the ‘within-
168 GCM’ variabilities of different scenarios; further analysis was done to evaluate the number of
169 simulations needed to represent the variability in Section 2.2.2.

170 **2.2 Rainfall data analysis**

171 **2.2.1 Rainfall characteristics**

172 Model simulations (*i.e.* historical and future rainfall projections) and BoM rain gauge data were
173 compared using a set of common rainfall characteristics at 6-min time steps. The purpose was to
174 better understand the behaviour of each GCM so that subsequent changes in WSUD reliability could
175 potentially be explained. Characteristics investigated included: *Annual Rainfall* (mm), *Number of*
176 *Rainfall Events per year* (a count of the number of events that are recorded; a threshold of a minimum
177 six hours were taken to separate two consecutive rainfall events), *Event Duration* (hours), *Average*
178 *Rainfall Intensity* (mm/hr; average rainfall intensity across all the rain events recorded - propagated
179 from 6-minute data), *Maximum Rainfall Intensity* (mm/hr; the average of the highest rainfall intensity
180 of all the rain events recorded), *Average Dry Period* (hrs; average length of dry periods between two
181 events), and *Annual Rainy Days* (a day with recorded total rainfall ≥ 1 mm).

182 Box plots were created for each of the selected rainfall characteristics to compare downscaled and rain
183 gauge data over the same periods. These not only provided an insight into the variability of scenarios
184 being modelled, but how accurately GCMs can reproduce historical rainfall. An analysis of trends
185 would also signal important findings about the GCM model-generated rainfall time series and future
186 climate predictions in general.

187 2.2.2 Capturing variability in model rainfall characteristics

188 A large ensemble of simulations from models is highly desirable to robustly account for uncertainties
189 in hydrological design (Raut *et al.*, 2018), *e.g.* in this study 100 simulations were generated from each
190 GCM. However, it is often not convenient or practical to use the entire rainfall simulation data-set
191 generated for WSUD modelling. It was therefore required to develop a pragmatic approach to the
192 modelling by using fewer simulations while still ensuring that the extracted sample is representative
193 of the original data. In this study, subsets of the 100 scenarios from each GCM (*i.e.* subsets of 5, 10,
194 25 and 50 scenarios) were randomly selected 10,000 times to generate large number of possible
195 scenario combinations. These subsets were statistically compared to the original 100 GCM
196 simulations. Each subset was compared against the original 100 scenarios: (i) to ascertain whether the
197 reduced sample was statistically representative of the original data (*i.e.* has similar median) and (ii)
198 was clearly derived from the same continuous distribution at the 5% significance level. This was done
199 using Wilcoxon rank-sum and two-sample Kolmogorov-Smirnov tests (Gibbons and Chakraborti,
200 2011).

201 2.3 Propagation of future rainfall projections to stormwater models

202 2.3.1 Stormwater treatment performance

203 The Model for Urban Stormwater Improvement Conceptualisation (MUSIC) (eWater, 2014) and a
204 third party software package, known as the DaCapo Design Curve Generator (Bach and Dotto, 2016),
205 were employed to evaluate the change in WSUD treatment performance (presented as design curves)
206 under future climate using the downscaled rainfall projections as input with the number of simulations
207 selected for use. MUSIC is a conceptual model that uses 6-minute continuous rainfall data to simulate
208 rainfall-runoff, pollution generation and treatment processes across a user-defined catchment and
209 WSUD treatment train, widely used in the Australian urban water industry to evaluate the treatment
210 performance of WSUD assets. As multiple scenarios and a large number of system designs were
211 required, DaCapo was used. The software systematically generates MUSIC simulations files though
212 varying the parameter values and batch runs them to produce data with which to plot WSUD system

213 performance curves (*i.e.* pollutant removal efficiency vs. system size).

214 In this study, two WSUD technologies – stormwater biofilters and wetlands were selected for
 215 investigation, as they are currently the most widely used WSUD systems for stormwater treatment
 216 (Hatt *et al.*, 2006), and their treatment performance, as well as flow mitigation capacities have been
 217 well reported (Gogate *et al.*, 2017). System sizes were varied from 0.01 – 5% of the catchment
 218 impervious area (taken here as 1 ha, 100% impervious) for biofilters and 0.01 – 10% for wetland
 219 systems. Two different biofilter designs with extended detention depths of 0.1m and 0.4m and
 220 wetland designs having water depths of 0.1m and 0.5m were tested. Other input parameters are
 221 summarised in Table 2. Four key indicators were simulated through MUSIC: flow reductions, load
 222 reductions of Total Suspended Solids (TSS), Total Phosphorus (TP) and Total Nitrogen (TN). Results
 223 were expressed in terms of these four indicators and plotted as performance curves against varying
 224 system sizes using both BoM rain gauge data and rainfall projections for all selected GCMs. It should
 225 be noted that biofilter and wetland systems were investigated individually in this study, *i.e.* one single
 226 WSUD technology in a catchment, to understand the specific impact of climate change on individual
 227 system.

228 **Table 2** Key design inputs used in MUSIC to model pollutant removal efficiency for biofilters and wetlands

System Type	Design Parameter(s)
Biofilter System	
System Surface Area [% catchment impervious area]	0.01 – 5 ⁽¹⁾
Extended Detention Depth [m]	0.1 and 0.4
Submerged Zone Depth [m]	0.4
Filter Depth [m]	0.5
Media Saturated Hydraulic Conductivity [mm/hr]	180
System Exfiltration Rate [mm/hr]	0.0 ⁽²⁾
Constructed Wetland System	
System Surface Area [% catchment impervious area]	0.01 – 10 ⁽¹⁾
Extended Detention Depth [m]	0.1 and 0.5
Permanent ponding depth [m]	0.35
System Exfiltration Rate [mm/hr]	0.0 ⁽²⁾
Detention Time [hrs]	72
Treatment Targets to be meet⁽³⁾:	
Total suspend solids (% load reduction)	80
Total Nitrogen (% load reduction)	45
Total Phosphorus (% load reduction)	45

229 ⁽¹⁾ 30 different surface areas were modelled, increasing exponentially from 0.01 to 5 (for biofilter) or 10 (for
 230 wetland); ⁽²⁾ Both WSUD systems were assumed to be lined, *i.e.* water in the systems was not allowed to
 231 exfiltrate into the groundwater; ⁽³⁾ according to by Best Practice Environment Management (BPEM) guideline
 232 in Victoria, Australia (VSC, 1999)

233 2.3.2 *Impact on flow frequency*

234 WSUD is known to reduce peak stormwater flows. Hence, in this study, the potential of biofilters and
235 wetlands to reduce flow frequency in urban areas was estimated, mainly in terms of managing the
236 frequent flood events (*e.g.* return periods up to 2 years); the 6-min time series flow data exported from
237 MUSIC were examined to identify all the peak daily discharges (excluding zero flows that resulting
238 from no rain) for partial frequency analysis (for detailed methods please refer to Pilgrim (2007)), for
239 the above investigated catchment. In addition to the urbanised catchment (no WSUD
240 implementation), two WSUD catchments were further explored: (1) biofilter sized as 1% of
241 catchment area with extended detention depth (EDD) of 0.1 m and (2) wetland sized as 3% of
242 catchment area with EDD of 0.5 m; both were typical biofilter/wetland design size in Melbourne. The
243 generated historical and future rainfall projections, as well as observed BoM rain gauge data were
244 used to generate the flow frequency curves to show the likelihood of daily peak flow changes.

245 2.3.3 *Storage-Behaviour analysis*

246 To assess changes in the reliability of alternative water supply through stormwater harvesting using
247 rainwater tanks under future climate, a hypothetical storage-behaviour analysis was undertaken with
248 continuous simulation of inflow, outflow and changes in storage volume of a simple rainwater
249 tank according to mass balance principles (Fewkes and Butler, 2000; Liaw and Tsai, 2004; Mitchell
250 *et al.*, 2008). The storage-behaviour model was simulated at a daily time step using aggregating 6-
251 minute GCM downscaled rainfall. Imteaz *et al.* (2011) and Mitchell *et al.* (2008) both demonstrated
252 that using a daily time step in the water balance model can sufficiently determine a realistic tank
253 volume.

254 For this analysis, the runoff coefficient of the roof catchment was set to 1.0 where the generated
255 runoff from a designated household roof area of 200 m² was assumed to divert completely to the
256 connected storage tank. A constant daily demand (D) of 0.43 m³/day was set according to the reported
257 daily water demand for in Melbourne (Coombes and Kuczera, 2003); tank sizes between 1 to 10 m³
258 were simulated to express results as a storage-reliability curve. Water demand, tank storage, spillage

259 (using a ‘yield after spill’ order Mitchell *et al.*, 2008) and supply were then calculated following the
 260 sequence of equations listed below (Eq. 1 to 4). On a particular day, if the water storage is greater
 261 than the tank capacity C , the excess water will spill over and the tank storage level at the end of the
 262 day is reset equal to C . The amount supplied is then equal to the demand, D , or limited to the volume
 263 remaining in the tank, S_i , depending on how much water is available.

$$V_i = R \times A \quad (1)$$

$$S_i = V_i + S_{i-1} \quad (2)$$

$$V_{spill} = S_i - C \quad \text{for } S_i > C \quad (3)$$

$$Su_i = \begin{cases} D_i \rightarrow S_i = S_i - D & \text{for } S_i > D \\ S_i \rightarrow S_i = 0 & \text{for } S_i < D \end{cases} \quad (4)$$

264 where, V_i is the harvested rainwater inflow on the i^{th} day (m^3), R is the daily rainfall (mm), A is the
 265 roof area (m^2), S_i is the water stored in the rainwater tank (m^3), S_{i-1} is the amount of water at the
 266 beginning of time step/day before (m^3), V_{spill} is the spilled amount of water from the tank (m^3), C is
 267 the capacity of the rainwater tank (m^3), Su_i is the amount of water supplied each day, and D_i is the
 268 daily rainwater demand (m^3).

269 Volumetric reliability is used to quantify the performance of the rainwater harvesting system and was
 270 calculated as:

$$271 \quad R_e = \frac{\sum Su}{\sum D} \times 100 \quad (5)$$

272 where, R_e is the volumetric reliability of the tank, D is the cumulated demand over the simulated 10-
 273 year period and Su is the total amount of water supplied in response to this demand. Reliability curves
 274 are developed using the output from Eq. 4 with reliability (R_e) plotted against storage volume, *i.e.* tank
 275 capacity (C) to visually discern the effect of future climate conditions on stormwater harvesting
 276 reliability.

277 3. Results & Discussion

278 3.1 Statistical analysis of rainfall data

279 Figure 1 compares different rainfall characteristics (*e.g.* annual rainfall, intensities, durations, rainy
280 days, etc.) across the different data sets and time periods. By comparing historical projections and
281 future projections, a drier future was simulated by all GCMs, evidently by less annual rainfall
282 (average of 579.4 mm of all GCMs), smaller number of rainfall events per year (93), as well as longer
283 dry periods (43 hours) compared to that of historical projections: 667.8 mm annual rainfall, 110
284 events, and 40 hours dry periods respectively. The results also indicate less rain days each year for
285 most of the GCMs, except for CMCC-CMS and MIRO-C5. In addition, the observed trends for both
286 the average and maximum rainfall intensities vary across different GCMs, with some simulated higher
287 extreme maximum intensities (*e.g.* ACCESS1-1, BCC-CMS1, CMCC-CMS and CNRM-CM5); this
288 have good agreement with previous studies that suggest higher rainfall intensities in the future (*e.g.*
289 Rosenberg *et al.*, 2010). Other GCMs, however, generated lower extreme intensities in the future.
290 Across all eight GCMs, MIRO-C5 is found to be the ‘wettest’ model (annual rainfall between 950 and
291 1000 mm) and GFDL-CM3 the ‘driest’ (annual rainfall between 500-650 mm). Within each GCMs,
292 the variation between 100 simulations is, however, small except for maximum intensity (which has a
293 similar range between GCMs). This indicates that variations between GCMs should be prioritised
294 rather than variations within a single GCM to understand future climate variability.

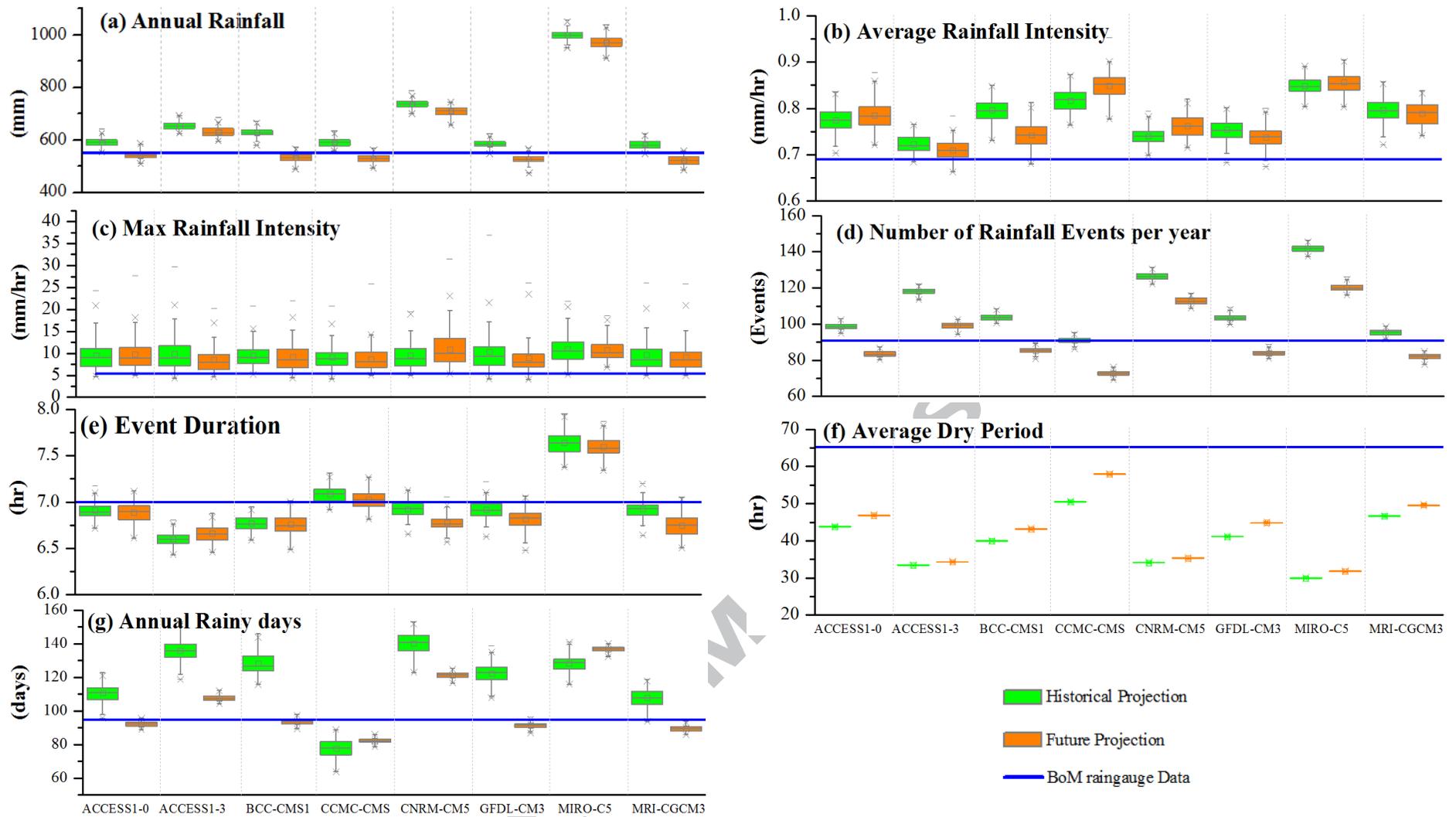
295 It was found that the rainfall characteristics of historical rainfall projections from the studied GCMs
296 did not have perfect match with that of observed BoM rain gauge data (Figure 1). Discrepancies
297 between rain gauge data and the projections from GCMs is evident, especially for *Average Dry*
298 *Period* (underestimated: 30-50 hours by GCM compared to 65 hours by BoM data; Figure 1); both
299 average and maximum rainfall intensities estimated from GCM projections were generally higher than
300 that from BoM data. Differences are expected as these GCMs are usually not calibrated directly to the
301 observed rainfalls but rather the atmospheric conditions to produce the global-scale atmospheric
302 processes (IPCC, 2013). Hence while GCMs may have good fit with historical atmospheric conditions,

303 the rainfall produced may be very different from historical measurements. The downscaling process
304 attempts to reflect the spatial variability within a single spatial domain. HiDRUS has been able to
305 reproduce spatial patterns of rainfall across Melbourne, Australia's domain as has been demonstrated
306 (Raut *et al.*, 2018). However, further distinguishing a single point (*i.e.* Melbourne Regional Office)
307 within this spatial domain for reasons of comparison with historical data will exacerbate uncertainties.
308 The authors believe that further calibration of the data against observations may actually risk altering
309 the model's intended behaviour entirely. As such, instead the uncertainties of each GCM were
310 embraced by exploring their 'within model' variability as well as their use as an ensemble future
311 rainfall data set. All selected GCMs are therefore included for further analysis; as it provides a more
312 robust assessment of the variability of WSUD reliability under future climate. Given the above
313 arguments further over-calibration of the downscaled time series are refrained, but the following
314 analysis of the results are performed while keeping in mind the actual differences between
315 downscaled historical GCM simulations and BoM observations.

316 ***3.2 Variability analysis of model simulations***

317 The frequency distribution curve in Figure 2 shows that a high level of similarity is retained when
318 reducing the sample size from the original 100 simulations downscaled from each GCM with even 5
319 simulations yielding significant proportions of p -values greater than 0.05. This is true for all models
320 as only slight variations are observed. It is therefore suggested that, for 10-year future scenario rainfall
321 time series, 5 to 10 simulations from a single GCM's downscaled outputs would be sufficient to
322 capture intrinsic model variability and maximise computational efficiency, also allowing modellers to
323 consider a greater variety of GCMs or climate scenarios in their assessments. These findings hold true
324 for Melbourne, but would have to be assessed for other locations. Findings may also differ for shorter
325 time series.

326



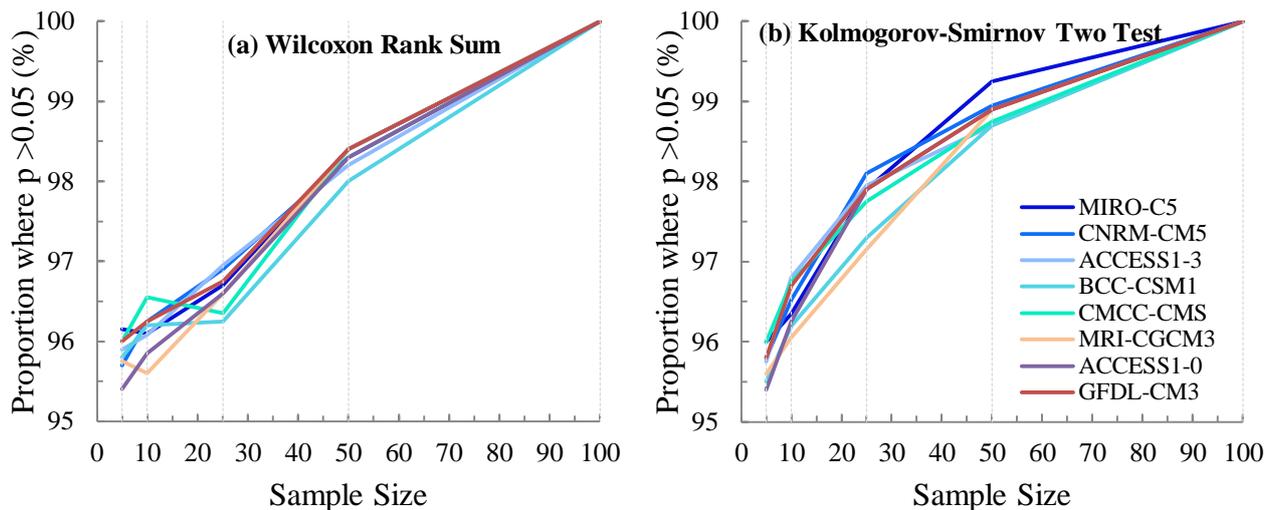
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Figure 1: Box plots of rainfall characteristics at Melbourne Regional Office using 6-minute ensemble of GCM simulations and rain gauge data over the baseline period of 1995-2004 (historical projections) and future period of 2040-2049 (future projections). 100 simulations from each GCM were used for this investigation.



331
 332 **Figure 2:** Frequency distribution curve for proportion of sample sizes (5, 10, 25, 50 and 100) that
 333 show significant similarity with the original 100 ensemble simulations for each GCM. GCMs are
 334 indicated with legends following a colour gradient from wettest (top dark blue: MIRO-C5) to driest
 335 (bottom orange: GFDL-CM3).

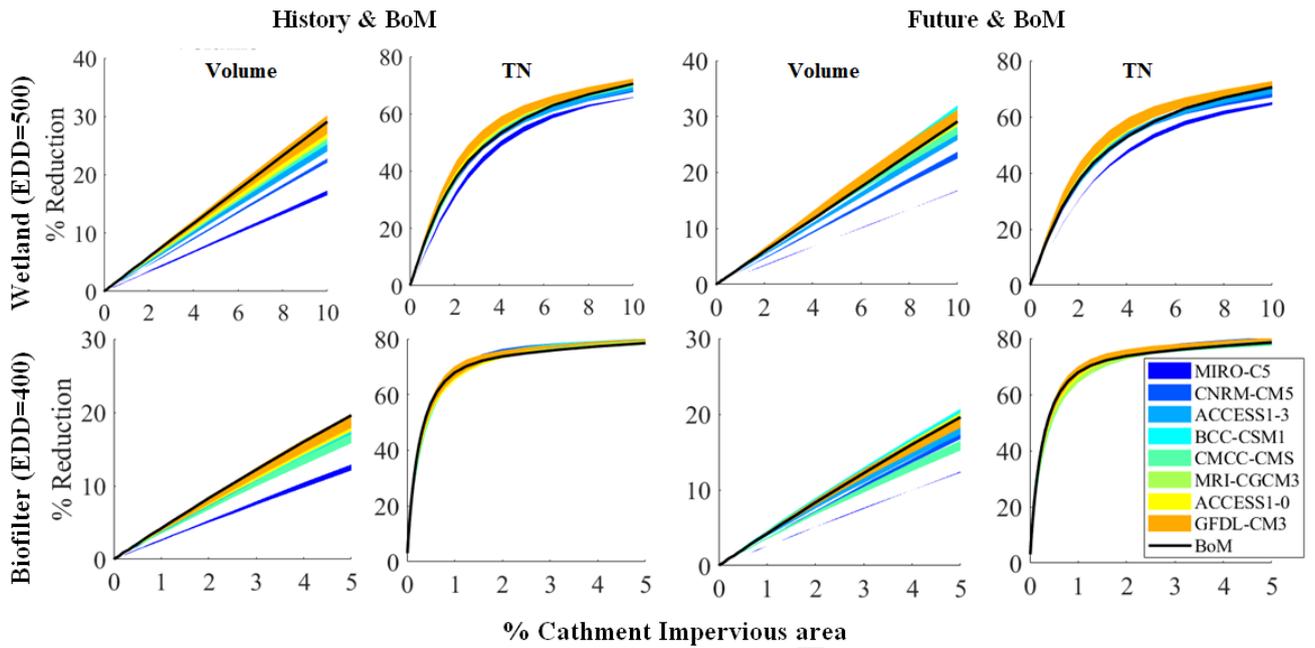
336 3.3 Stormwater treatment performance

337 Figure 3 presents the design curves generated using historical and future rainfall projections, with the
 338 comparison to one generated based on BoM observation data (only volume and TN curves are
 339 presented here as examples due to the high similarities between the pollutants). Interestingly, despite
 340 the statistical differences of rainfall characteristics identified in Figure 1 between historical and future
 341 rainfall projections, the design curves in Figure 3 show high similarities between the historical and
 342 future ones. However, slightly higher runoff volume reduction (~3%) is simulated from future rainfall
 343 projections compared to historical conditions (Figure 3). This finding is further confirmed by the
 344 direct comparison of WSUD treatment performance estimated from historical and future rainfall
 345 projections (Figure 4) — all points are closely centred around the 1:1 “no-impact” line (*i.e.* future
 346 equals to history) for pollution reduction, while majority of the points (>95%) for runoff volume
 347 reduction are above the “no-impact” line. The increased performance of runoff volume reduction in
 348 the simulated future is likely due to decreased inflow volumes in the drier future, which was found
 349 previously to lead to higher volume reductions (Hatt *et al.*, 2009); pollutant removal in WSUD
 350 systems, however, is influenced by multiple factors *e.g.* it was reported that longer dry periods can

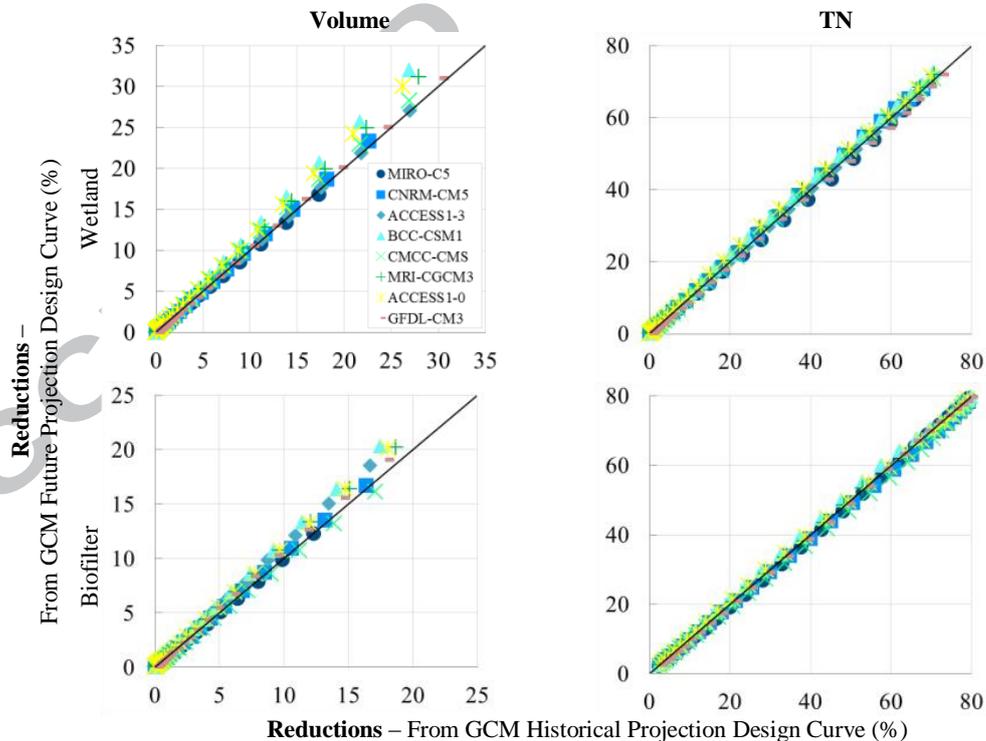
351 adversely impact system performance in removing sediments and nutrients in stormwater (Hatt *et al.*,
352 2009; Zhang *et al.*, 2015). Nevertheless, other design elements/operational factors of biofilters (*e.g.*
353 plants, submerged zones and maintenance), which are not accounted in MUSIC may become crucial
354 when considering future climate. For example, a notable increase in dry period or decrease in the
355 number of rainy days (see Figure 1 f and g) will require a change in design of how water is retained in
356 the system or more frequent maintenance to ensure that plants within the system are healthy.

357 Figure 3 also indicates variability among different scenarios within a single GCM and different GCMs.
358 Bigger variations of reductions are observed between GCM predictions than within a GCM. GCM
359 projections produce broader bands for wetland systems compared to biofilters, showing that wetlands
360 appear more susceptible to future climate. This could also be due to of larger system size and scale,
361 *e.g.* wetlands are usually an order of magnitude larger than biofilters.

362 Table 3 summarises the desired system size to meet treatment targets required by BPEM guideline in
363 Victoria, Australia (Table 2) estimated using the BoM curve as well as the GCM curves in Figure 3.
364 According to the estimates, in general, large uncertainties exist considering various historical and
365 future scenarios with the latter ones providing higher variability. It can be seen that if wetlands are to
366 be designed to achieve BPEM targets, they should be sized to 3.5% of catchment impervious area
367 according to the BoM curve (which currently underpins system design and compliance checking in
368 practice); depending on different future scenarios, system sizes can vary from 2.5% to 4.0% of the
369 catchment area, indicating that the current design may be sufficient but under some GCM scenarios
370 (*e.g.* MIRO-C5), larger systems may be required. In terms of biofilters, the current designed system
371 size (0.8% of the catchment impervious area) is at the lower boundary of the variability bands
372 estimated from all GCMs (0.8-1.5% of catchment area; Table 3) to deliver the same treatment
373 performance. While it is often not necessary, biofilters could be sized to 1.5% of the catchment area
374 for ensuring 100% reliability under future conditions according to the current analysis. MICRO-C5,
375 which is the 'wettest' GCM, always guides the critical system design for both systems.



376
 377 **Figure 3:** Wetland (EDD=500mm) and biofilter (EDD=400mm) design curves. The shaded bands
 378 indicate the variation between the 5 simulations used in each GCM. GCMs are indicated with legends
 379 following a colour gradient from wettest (dark blue: MIRO-C5) to driest (orange: GFDL-CM3).



380 **Figure 4:** Comparison of treatment performance estimated using historical design curves and future
 381 curves for wetland (EDD= 500mm) and biofilter (EDD = 400 mm).

382 **Table 3** *Estimated sizes of the systems required to meet BPEM targets (i.e. 80% TSS, 45% TP and*
 383 *45% TN load reduction) using historical and future projections, as well as BoM rainfall*

	Wetland ¹⁾ (Historical)	Wetland (Future)	Biofilter (Historical)	Biofilter (Future)
MIRO-C5	3.8-4.0% ³⁾	3.7-4.0%	1.4%	1.4-1.5%
CNRM-CM5	3.3-3.4%	3.1-3.5%	1.1-1.2%	1.1-1.2%
ACCESS1-3	2.9-3.5%	3.2-3.6%	1.0-1.1%	0.9-1.1%
BCC-CSM1	2.8-3.2%	2.8-3.2%	0.9-1.0%	0.9-1.0%
CMCC-CMS	3.1-3.2%	3.2-3.4%	0.9-1.0%	0.9-1.0%
MRI-CGCM3	3.1-3.2%	2.9-3.1%	0.9-1.0%	0.9-1.0%
ACCESS1-0	2.8-3.1%	2.7-3.2%	1.0-1.1%	0.9-1.1%
GFDL-CM3	2.5-3.1%	2.5-3.1%	0.8-1.0%	0.8-0.9%
BoM	3.5%	3.5%	0.8%	0.8%

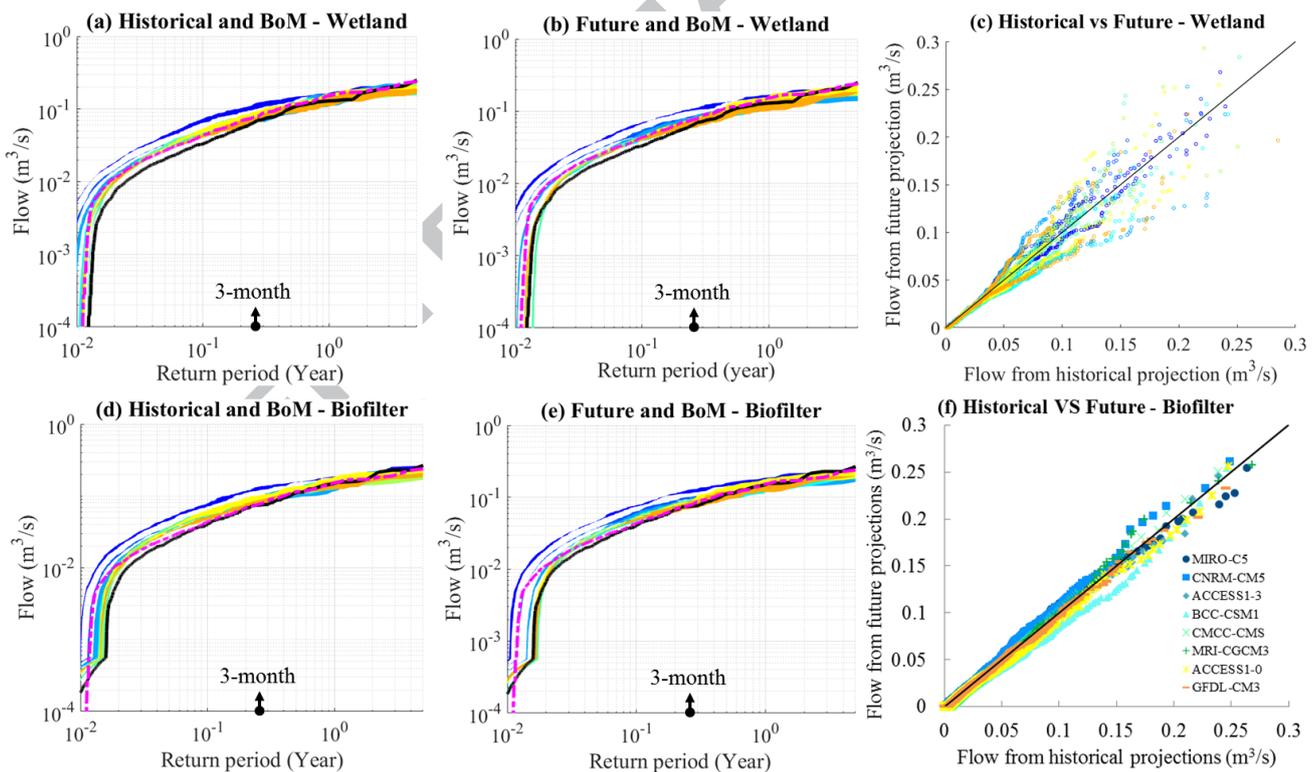
384 *Note: ¹⁾ the critical pollutant for wetland design is TSS; ²⁾ the critical pollutant for wetland design is TP; ³⁾ sizes are*
 385 *expressed as percentage of catchment area;*

386 **3.4 Impact on flow frequency**

387 The comparisons of peak flows estimated by MUSIC using historical and future rainfall projections,
 388 as well as observed BoM rainfall are presented in Figure 5. As indicated, in general climate change
 389 simulated through HiDRUS from different GCMs shows high similarity of flows between historical
 390 and future scenarios and with the direct comparisons in Figure 5c&f show limited impact for small
 391 flows (*i.e.* <0.1 m³/s) with low variability, closely aligning to the 1:1 “*no-impact*” line. This explains
 392 why minimal differences of pollution removals were estimated between historical and future
 393 projections found in Figure 3 as WSUD systems are usually designed for small events with flows up
 394 to 3-month return period (=0.25 [Years], peak flow is equivalent to approximately 0.1 m³/s) for
 395 pollution treatment. While for higher flows (>3-month return period), points scatter further away from
 396 the “*no-impact*” line and the differences vary according to different GCMs, indicating higher
 397 variabilities between GCMs. For example, the GCMs with both higher average and maximum rainfall
 398 intensities in the future projections (*e.g.* CNRM-CM5, Figure 1) often have higher estimated flows
 399 than the ones with lower estimated average and maximum rainfall intensities (*e.g.* BCC-CMS1).

400 The flow frequency behaviour of a fully urbanised catchment with no WSUD systems implementation
 401 (estimated from the BoM rainfall by MUSIC) was also plotted in Figure 5. Biofilters or wetlands can
 402 provide the greatest benefit up to a return period of around 3-month, after which there are minimal
 403 differences with that of an urbanised catchment in terms of peak flows, potentially indicating the

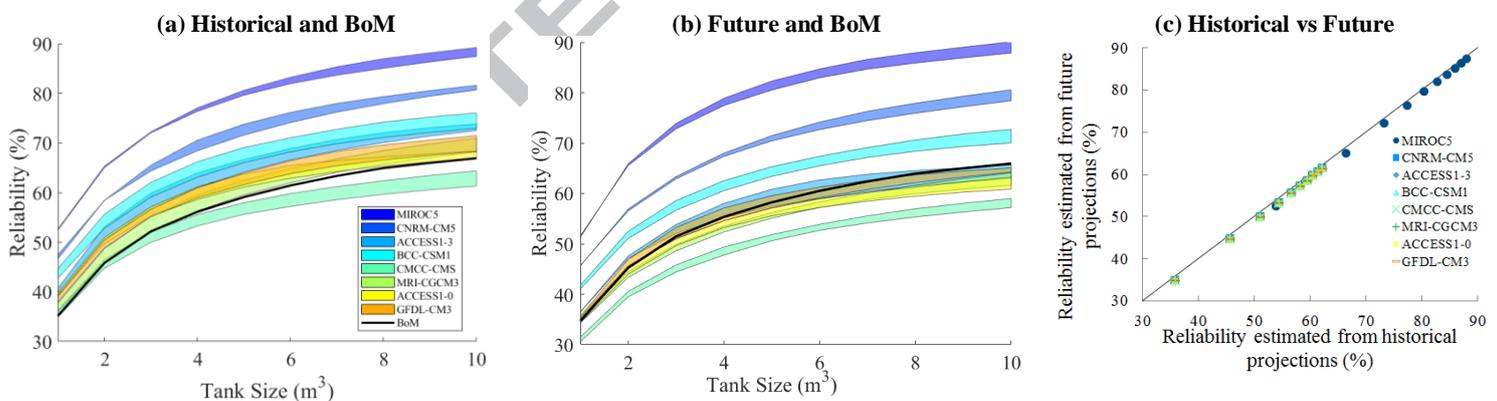
404 ineffectiveness of these systems under such significant events; biofilters and wetlands are able to
 405 reduce the peak flow of 3-month return period rainfall event from $0.085\text{m}^3/\text{s}$ to $0.075\text{m}^3/\text{s}$ and
 406 $0.070\text{m}^3/\text{s}$, respectively. Again, considering future climate, large uncertainties are observed between
 407 GCMs, and even with WSUD implementation, the estimated flows often exceed the values of the
 408 urbanised catchment at present day. The estimated 3-month flows vary from $0.06 - 0.12 \text{m}^3/\text{s}$ for the
 409 WSUD catchments and the wettest model (MIRO-C5) represented the most critical design scenario
 410 (*i.e.* = $0.12 \text{m}^3/\text{s}$); this is expected as MIRO-C5 has the highest intensities and longest event durations
 411 (Figure 1). This also explains why the BoM curves are closer to the lower bounds of the curves as
 412 BoM rainfall data has relatively lower rainfall intensities. Comparative evaluation also illustrates
 413 another risk of selecting a singular “best” model as if selecting wettest GCM MIRO-C5 might lead to
 414 overly conservative predictions of larger, more frequent flows.



415
 416 **Figure 5:** Comparisons of daily peak flows generated from historical and future projections, as well
 417 as observed BoM rainfall; Year in log used in X-axis; ‘BoM (Urbanised – No WSUD)’ refers to the
 418 with no biofilter or wetland systems).

419 3.5 Stormwater harvesting reliability

420 Figure 6 shows the storage-behaviour analysis results (plotted as reliability against tank size)
 421 estimated using both historical and future projections, and BoM rainfall. Different from treatment
 422 performance and flow reduction that have high similarity between historical and future conditions,
 423 simulated climate change always lead to slightly lower harvesting reliabilities (<3%) for all GCMs
 424 (see Figure 6), very likely due to drier conditions in the future. A clear difference in harvesting
 425 reliability between GCMs is observed, with higher reliability estimated for wetter GCMs. For
 426 example, for a 3 m³ rainwater tank, the estimated reliability in the future can vary from 45% (CMCC-
 427 CMS) to 72% (MIRO-C5), indicating significant future uncertainties for harvesting reliability through
 428 rainwater tanks. The variation between simulations within each GCM is still minor (e.g. <3% for 10
 429 m³ tank) and large variation evidenced across GCMs. The resulting critical scenario for this case
 430 (CMCC-CMS) is different to that for treatment performance and flow frequency impact where the
 431 MIRO-C5 is the critical GCM (Figure 3; Figure 6). This may be attributed to the lower rainfall,
 432 greater dry period and less rainy days recorded for CMCC-CMS (Figure 1).



433 **Figure 6:** Storage behaviour curves illustrating the relationship between rainwater tank size and
 434 reliability under the different climate scenarios.

435 4. Practical implications and limitations of the work

436 Adapting to climate change will require rejecting basic assumptions about stationary conditions that
 437 have historically underpinned flood, water, and conservation management (Milly *et al.*, 2008). Some
 438 argue that simply coping with present climate variability is enough of a challenge (Washington *et al.*,

439 2006), however, with the understanding that infrastructure is generally less costly and disruptive if
440 necessary measures to mitigate climate change are taken well in advance of anticipated changes.

441 Given the complexity of global warming and climate uncertainty, which is well exemplified by
442 distinctive statistical characteristics of each GCM in Figure 1, it is necessary to look beyond the
443 notion of a singular “best model” and towards the use of an ensemble in order to provide assurance
444 and robust assessment of system reliability under a wide range of potential conditions. The fact that
445 there was no unambiguously superior model observed (*i.e.* downscaled historical GCM projections
446 did not match perfectly with observations) and, most importantly, that the critical design scenario
447 varied according to the performance indicators (*i.e.* treatment vs harvesting) supports this hypotheses
448 that it is better to use less simulations and more models than less models and more simulations. It was
449 found that using a large number of realisations from each downscaled GCM is not necessary if
450 designers wish to most efficiently and comprehensively assess the reliability of their systems over the
451 long term. In this study only 5 realisations were used, and the actual number for different cases shall
452 be determined according to specific rainfall data, the study area and the objective criteria.

453 The results in this paper suggest that the WSUD performance in pollutant removal and flow
454 reductions under simulated future climates has minimal difference to that under simulated historical
455 conditions. Stormwater harvesting systems are expected to have slightly lower reliabilities (<3%)
456 under future climate. Significant uncertainties exist according to future rainfall projections across
457 various GCMs; looking into these uncertainties can provide insights to the performance variabilities,
458 and assisting the adaptation of WSUD systems into unknown future. The analysis reveals that while
459 WSUD systems may stay resilient in providing treatment performance under the simulated future,
460 they could be also sized bigger to account for future climate uncertainties. Notably, this analysis only
461 considered impact of future rainfall in treatment performance, while in fact the stormwater pollution
462 concentration may increase due to urbanisation (Wang *et al.*, 2017), hence further studies are
463 recommended to understand the impact of urbanisation on WSUD reliability. Moreover, peak flows
464 (especially for return periods of < 3 months) based on BoM rainfall are often underestimated
465 compared to the simulated future climate variability; this was probably due to the higher estimated

466 rainfall intensities from both historical and future projections compared to BoM observations. Larger
467 rainwater tank sizes may be needed to provide same level of harvesting reliability considering future
468 climate change scenarios projected from any GCM, *e.g.* to achieve 50% of reliability, 3 m³ tank size
469 is typically enough at present according to BoM rainfall data, it however needs to be increased to at
470 least 4.5 m³ to keep the same reliability under the future scenarios based on the CMCC-CMS (Figure
471 6).

472 It was noted that there were differences between downscaled historical data and observations, which
473 may potentially influence the results (*e.g.* the underestimation of peak flows from BoM rainfall).
474 Moreover, this study was limited to the results of just eight climate model projections at a single
475 location, with only one CO₂ emission scenario – RCP8.5 scenario from CMIP5 that represents no
476 action in climate change mitigation. Further, the uncertainties in the models used have not been
477 discussed. Therefore the findings from the current study should be taken as preliminary, future
478 research should be directed towards replicating and expanding this study and reproducing these results
479 for a larger database with more locations and models to further justify these claims. Also, studies can
480 also be expanded to involve more WSUD technologies, *e.g.* ponds and swales; nevertheless, all these
481 analysis indicate the necessity of considering future uncertainties when designing stormwater
482 management systems; and the method in this study can be used to quantitatively design a conservative
483 WSUD system to cope with future climate variability.

484 **5. Conclusion**

485 In this paper, a multiplicative cascade model (HiDRUS) was employed to generate high resolution
486 rainfall projections at 1 km and 6 minute interval for an ensemble of General Circulation Models
487 (GCMs) at Melbourne Regional Office to offer, for the first time, comprehensive insights into the
488 adaptability of Water Sensitive Urban Design (WSUD) under a variety of future climate scenarios on
489 a much fine scale. Four critical contributions and conclusions from this study were found:

490 1. Downscaled rainfall projections from eight different GCM's have been statistically

- 491 characterised; the results indicate the likelihood of a drier future with less rainfall events and
492 longer dry periods;
- 493 2. Despite their validity at the atmospheric level, rainfall projections do not necessary match the
494 observed rain gauge data. Nevertheless, these discrepancies and uncertainties associated with
495 their predictions should be embraced and propagated to WSUD reliability rather than discarded;
- 496 3. Using an ensemble of GCMs is preferable over a “best model, many scenarios” approach to
497 account for the variety of potential scenarios; it is found that for this location and setup only 5
498 rainfall simulation scenarios (10 years in length) are needed to capture the variability of the
499 downscaled rainfall simulations from each tested GCM;
- 500 4. Minor differences of WSUD performance in removing pollutants, flow frequency reduction,
501 and slightly lower harvesting reliability were observed when comparing simulated historical
502 conditions with future scenarios; however high variabilities do exist and larger WSUD systems
503 are suggested to cope with the high variabilities in the simulated future scenarios, in order to
504 ensure that the treatment targets as well as harvesting reliability are still met.

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621 **Highlights**

- 622 • High resolution catchment scale rainfall predictions were generated from 8 GCMs
- 623 • Multiple GCMs is preferable over a best model to account for the future variability
- 624 • Simulated climate change has limited impact on pollutant treatment performance
- 625 • Larger WSUD systems are recommended to account for the future variability
- 626 • Bigger rainwater tank is suggested for same harvesting reliability in the future

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ACCEPTED MANUSCRIPT