



The effect of environmental stochasticity on species richness in neutral communities



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HIGHLIGHTS

- The recently proposed time-averaged neutral theory is analyzed.
- The effect of demographic stochasticity on the storage mechanism is examined.
- Unimodal effect of environmental stochasticity on species richness is explained.

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ABSTRACT

Environmental stochasticity is known to be a destabilizing factor, increasing abundance fluctuations and extinction rates of populations. However, the stability of a community may benefit from the differential response of species to environmental variations due to the storage effect. This paper provides a systematic and comprehensive discussion of these two contradicting tendencies, using the metacommunity version of the recently proposed time-average neutral model of biodiversity which incorporates environmental stochasticity and demographic noise and allows for extinction and speciation. We show that the incorporation of demographic noise into the model is essential to its applicability, yielding realistic behavior of the system when fitness variations are relatively weak. The dependence of species richness on the strength of environmental stochasticity changes sign when the correlation time of the environmental variations increases. This transition marks the point at which the storage effect no longer succeeds in stabilizing the community.

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1. Introduction

One of the biggest puzzles in community ecology is the persistence of high-diversity assemblages. The competitive exclusion principle (Gause, 2003; Hardin et al., 1960) predicts that the number of species coexisting in a local community should be fewer than or equal to the number of limiting resources, in apparent contrast with the dozens and hundreds of locally coexisting species of freshwater plankton (Hutchinson, 1961; Stomp et al., 2011), trees in tropical forests (Ter Steege et al., 2013) and coral reef species (Connolly et al., 2014). This problem has received considerable attention in recent decades, with many mechanisms suggested to circumvent the mathematical constraints embodied in the exclusion principle and many works that try to provide empirical support to these theories (Chesson, 2000; Gravel et al.,

2011; Amarasekare, 2003).

Within this framework, neutral theories, and in particular the neutral theory of biodiversity (NTB) suggested by Hubbell (2001), Volkov et al. (2003), and Rosindell et al. (2011), play an important role. Under neutral dynamics all individuals are considered as having the same fitness, and abundance variations are the result of demographic noise alone. The number of individuals belonging to each species varies randomly within the limit imposed by the overall size of the community, with most populations eventually drifting to extinction. However, the neutral turnover rate is very slow, and diversity is maintained due to the introduction of new species into the system, either via speciation (in the metacommunity) or by migration (in a local community).

The slow turnover dynamics in the neutral model is not only an advantage, it is also a disadvantage, and has triggered one of the main lines of criticism directed at the neutral model. It turns out that pure ecological drift is far too slow to account for both the observed short-term fluctuations and the long term dynamics (Ricklefs, 2006; Nee, 2005; Leigh, 2007; Kalyuzhny et al., 2014a,

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2014b; Chisholm et al., 2014). For example, the abundance of the most common species in the Barro-Colorado Island Smithsonian 50 ha plot has decreased from 40,000 to 30,000 individuals (>1 cm dbh) during about half of a generation, while under pure demographic noise one expects variations of order $\sqrt{N} \sim 200$ within a whole generation. The abundance of the most common species in the Amazon basin is about 10^9 individuals (Ter Steege et al., 2013). Under neutral dynamics, this is the expected age (in generations) of that species, and since the generation time for tropical trees is about 50 years, this timescale ($5 \cdot 10^{10}$ y) is longer than the age of the universe (Ricklefs, 2006; Nee, 2005; Leigh, 2007). Recent work (Chisholm and ODwyer, 2014; see also Danino and Shnerb, 2015) shows that species' ages in neutral models are in fact lower than these early estimates by about two orders of magnitude, however these ages are still too high to be realistic.

Motivated by these difficulties, recent works (Kalyuzhny et al., 2015; Kessler and Shnerb, 2014) have pointed towards a generalized neutral theory that will include both demographic and environmental stochasticity. Basically, this new model accepts the equivalence principle, but assumes that the fitness of all species, when averaged over time, are equal while at any instant some species have higher fitness than the others due to temporal variations in parameters such as temperature, precipitation, etc. Accordingly, all species are equivalent and abundance variations are driven by fluctuations (Azade et al., 2016). The ability of this time-averaged neutral theory of biodiversity (TNTB) to explain various empirical patterns, including species abundance distributions, temporal fluctuations statistics and the growth in system dissimilarity over time, was demonstrated in Kalyuzhny et al. (2015).

However, by introducing a species-specific response to environmental variations, the TNTB finds itself entering the domain of another celebrated mechanism that was suggested to explain species coexistence, the storage effect introduced by Chesson in the 1980s. In particular, Chesson and Warner (1981) considered the "lottery game" in which the fitness of each species, as reflected by the chance of its offspring successfully occupying a vacancy in the community, fluctuates in time. This differential response of species, when superimposed on buffered population growth and covariance between environment relative probability and competition (Chesson, 1994) was shown to stabilize the system. Chesson and Warner showed how rare species, when compared with common species, have fewer per-capita losses when their fitness is low and more gains when their fitness is high. Accordingly, the population of rare species increases (their average growth rate is positive just because their relative abundance is low) and the system supports a stable equilibrium: species' abundance fluctuates, but are attracted to a finite value by a restoring force.

Hubbell's NTB, which takes into account demographic noise and speciation but with no environmental noise, provides us with one set of predictions for the patterns characterizing a community, such as species abundance distribution and species richness. The Chesson–Warner lottery game, taking into account only environmental stochasticity (without demographic noise or speciation) suggests another set. What happens under the general model of TNTB, where *all* these elements play a role? What patterns does it predict, and how do they depend on the strength of the storage effect? In Kalyuzhny et al. (2015) the TNTB was presented in the context of a mainland-island model and simulated island dynamics were compared with data from the Barro-Colorado Island (BCI) plot. Here we aim at understanding the metacommunity dynamics of the TNTB and to explore its relationships with both NTB and the lottery game.

To do that, we first revisit the storage effect, using the original Chesson–Warner model. In Section 2 we consider the storage

effect for two species, emphasizing the transition it shows from a balanced system, where the abundance of both species fluctuates around one half of the community, and an imbalanced state, with one rare and one frequent species. A deeper analysis of the equilibrium distribution poses a conceptual problem, namely that the result is independent of the amplitude of the environmental variations. This problem is discussed in Section 3, and indicates the necessity of incorporating demographic stochasticity into the model. Before doing that, in Section 4 we consider the original lottery game for communities with many species and discuss its applicability to empirical systems. Finally in Section 5 the TNTB model, in which environmental variations, demographic stochasticity and speciation affect the community, is analyzed. Conclusions are presented in the last section.

2. A lottery game for two species

In this section we study the simplest case, the storage effect in a community with two species playing the lottery game. Since we are ultimately interested in the TNTB, we assume that the fitness of both species is equal when averaged over time (species are equivalent). Note that the scope of the storage effect is wider, and it may stabilize a community even when the average fitnesses are different; we will return to this point in the discussion section.

To provide an intuitive numerical example, let us consider an extremely simple game. Imagine a forest with 100 trees, N_A of species A and $N_B = 100 - N_A$ of species B. For simplicity we assume that there is no spatial structure, seeds and seedlings of both species are all around the forest, with relative frequencies that reflect the relative abundance of adult trees. During every year 20% of the trees are selected at random, independent of their species affiliation, to die (so that the generation time is five years). The gaps that remain after the trees' death are filled by seedlings, where the chance of each seedling to capture the vacancy depends of its relative fitness, with the fitness varying in time. To have equivalent species the temporal fitness is taken to be an independent and identically distributed (i.i.d.) variable, so the chance of a particular species to be the fitter of the two in a certain year is 1/2. Under an extreme, "winner takes all" scenario, the fittest species of a given year captures all the 20 empty slots.

Now let us follow the dynamics. Consider the case where, at the beginning of a certain year, $N_A = 20$ and $N_B = 80$. After the death step, $N_A = 16$ and $N_B = 64$ (this is an average, since trees are picked to die at random, but for our purpose it is sufficient to trace the average). Now there are two options: if the winner of this year is species A, the year ends with $N_A = 36$, $N_B = 64$, while if the fittest species is B, the outcome will be $N_A = 16$, $N_B = 84$. One can easily see that the gain of A when it wins, 16, is higher than the potential gain of B, which increases its population only by four individuals when it wins. By the same token the losses of A when it is the inferior species are smaller than the losses of B in the parallel situation.

While this example is misleading in several respects (in particular the unrealistic winner takes all assumption strongly affects the results), it still provides the basic intuition: although the average fitness of both species is the same, environmental variations provide benefit to the rarer one, as the opportunities for the rare species (when it wins) are greater than those of a common species and its risks (when it loses) are less. Accordingly, an effective stabilizing force acts against any deviation from the 50–50 partition.

Having established this intuition, let us turn to the original two-species model as presented in Chesson and Warner (1981). In this model there is no demographic noise, so the absolute number of individuals has no importance. Accordingly, the variables are

species relative fractions. For two species, these are x_1 and $x_2 = 1 - x_1$.

The model has a two step dynamics. During the death step a fraction δ of the trees are removed, so the loss of species number 1, for example, is δx_1 . The gaps are filled by seedlings. The number of seedlings for a species is proportional to its abundance, and the chance of a single seedling to capture the empty slot is determined by its species' fitness. Accordingly the abundance of the two species after these steps, death and recruitment, is given by,

$$\begin{aligned} x_1^{t+1} &= x_1^t(1 - \delta) + \delta \frac{f_1^t x_1^t}{f_1^t x_1^t + f_2^t x_2^t} \\ x_2^{t+1} &= x_2^t(1 - \delta) + \delta \frac{f_2^t x_2^t}{f_1^t x_1^t + f_2^t x_2^t} \end{aligned}$$

where $f_i^t > 0$ is the fitness of the i -th species during the t -th step. For our two species system one can replace x_2 by $1 - x_1$ to get a single recursion equation,

$$x_1^{t+1} = x_1^t(1 - \delta) + \frac{\delta f_1^t x_1^t}{f_1^t x_1^t + f_2^t(1 - x_1^t)}. \quad (1)$$

When the fitness is fixed in time, the fittest species will win the game and the abundance of the inferior species decreases monotonically towards zero. Chesson discovered that, when the fitness fluctuates in time, it stabilizes the populations. As we are interesting in time-averaged neutral models, we require the long-term averages of f_1 and f_2 to be equal.

Two parameters are needed to characterize environmental stochasticity: its strength and its duration (correlation time).

1. The *strength* of the environmental stochasticity, σ_E^2 , manifests itself in the spread of the fitness parameters f_i . Without loss of generality one may take

$$f_i = e^{\gamma_i}, \quad (2)$$

where the parameter γ_i is an independent and identically distributed (iid) variable picked from a distribution (say, a Gaussian or a uniform distribution) with zero mean and variance σ_E^2 . If $\sigma_E^2 = 0$ then $f_i \equiv 1$, all species have the same fitness and the dynamics stops, $x_i^{t+1} = x_i^t$. The larger σ_E^2 , the stronger the fitness variability.

2. δ is the correlation time of the environmental noise, measured in units of generations (see also the description of our agent-based simulation of the model in Appendix B). Our analysis of Eq. (1) is based on the assumption that the fitness f_i is picked at random every elementary timestep, i.e., between t and $t + 1$ and so on. Within this period a fraction δ of the individuals die. To give a concrete example, in Kalyuzhny et al. (2015) the correlation time of the environmental stochasticity in the Barro Colorado Island plot was found to be about 10 years, while the generation time is about 50 years. To model this dynamics using Eq. (1) one may take $\delta = 1/5$, meaning that the replacement of 1/5 of the trees takes place under (more or less) the same fitness regime.

Hatfield and Chesson (1989) showed how to map the discrete time equation (1) to a Fokker–Planck equation for $P(x_1, t)$, the probability that the relative abundance of species number 1 is x_1 ,

$$\begin{aligned} \frac{\partial P(x_1, t)}{\partial t} &= \delta \sigma_E^2 \left\{ \frac{\partial}{\partial x_1} \left[x_1(1 - x_1)(x_1 - 1/2)P(x_1, t) \right] \right. \\ &\quad \left. + \delta \frac{\partial^2}{\partial x_1^2} \left[x_1^2(1 - x_1)^2 P(x_1, t) \right] \right\}. \end{aligned} \quad (3)$$

We recapitulate the derivation of (3) from (1) in Appendix A.

The steady-state solution can be seen to be

$$P_{eq}(x_1) = C[x_1(1 - x_1)]^{\frac{1}{\delta} - 2}, \quad (4)$$

where $C = \Gamma(2/\delta - 2)/\Gamma^2(1/\delta - 1)$.

This result emphasizes two general features of the storage effect. First, the right hand side of the Fokker–Planck equation (3) has two terms. The first is the “drift” term, describing the dynamics of the average value of x_1 , which drives x_1 towards $1/2$. The “diffusive” term, involving the second derivative with respect to x_1 , has the coefficient $[x_1(1 - x_1)]^2$, meaning that the random wandering of the system is strongest when $x_1 = 1/2$ and approaches zero at the edges, $x_1 = 0$ and $x_1 = 1$ since there is no dynamics (and no fluctuations) if either species reaches fixation. As discussed in Ohkubo et al. (2008), the resulting $P_{eq}(x_1)$ reflects the balance between these two opposing forces: the diffusive aspects of the dynamics acts to trap the system close to the edges where the “diffusion constant” associated with abundance fluctuations vanishes while the drift term pushes x_1 to the stable fixed point in the middle.

The net result is determined by the ratio between these terms, i.e., by δ , as illustrated in Fig. 1: for $\delta < 1/2$ the deterministic term wins, leading to a distribution with a single maximum at $1/2$, meaning that at any instant of time the community is likely to be well balanced, with both species represented by roughly the same number of individuals. For $\delta > 1/2$, on the other hand, P_{eq} is convex, with probability mass concentrated near the edges at zero and one. In this case the community is unbalanced, (almost) any snapshot picture of the community reveals strong dominance of one species, although the equivalence ensures that the time average fraction of each species is around $1/2$.

The distribution peak for $\delta < 1/2$ resembles the Gaussian or exponential peak one finds when the system supports a deterministic stable equilibrium (an attractive fixed point of a nonlinear system), but this similarity is slightly misleading. The decay of P_{eq} towards the edges is described by a power-law, not by an

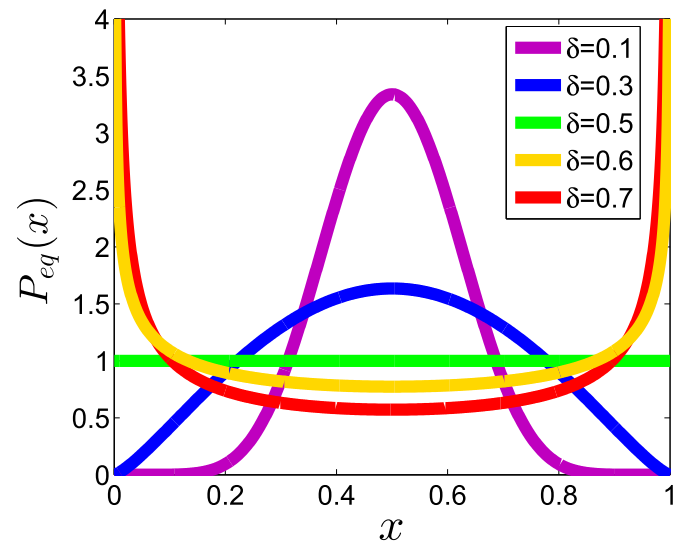


Fig. 1. The equilibrium probability distribution function, $P_{eq}(x)$, given by Eq. (4), is plotted against x for different values of δ (see legend). For $\delta < 1/2$ the distribution peaks around the symmetric point 0.5 , and the peak becomes sharper when δ decreases (still the decay is slower than exponential). At $\delta_c = 1/2$ the distribution is flat, and for smaller values of δ it develops two peaks close to the extinction and the fixation points and a valley in the middle. The distribution is normalizable as long as $\delta < 1$. Had the dynamics of Eq. (1) allowed an absorbing state (e.g., if one consider the fraction x smaller than x_{min} as the state with no individual) the chance of extinction in case of $\delta \ll 1/2$ would have been much smaller than the chance if $\delta > 1/2$. As discussed in the main text, the pdf's shown here are independent of σ_E^2 .

exponential or a Gaussian. This happens since the fixed point at $1/2$ is noise-induced in the first place.

The second key feature of Eq. (4) is that it has been derived from Eq. (1) by expanding it to the leading order in the fitness differences, so the emerging P_{eq} is independent of σ_E^2 (see Appendix A for more details). This approximation becomes better and better as σ_E^2 decreases; accordingly, the storage effect appears to stabilize the two-species community even for vanishingly small values of σ_E^2 . The amplitude of fitness variations only sets the time scale, such that the time needed for the system to reach the equilibrium distribution scales like $1/\sigma_E^2$ and diverges when the environmental noise vanishes, but P_{eq} itself stays the same. In the next section we discuss the conceptual difficulties associated with this outcome.

For the sake of completeness we note that the storage effect acts to destabilize the system when the environmental stochasticity acts on mortality rather than on fecundity. One may realize that easily by repeating the “loser loses all” version of the numerical example above, where the low fitness species suffers all the mortalities but the recruitment depends on abundance and is independent of fitness. Under such scenario the rarer species has more to lose than the commoner one, as a the fixed number of deaths would represent a larger share of its population (Hatfield and Chesson, 1989).

3. Storage effect and demographic stochasticity: a conceptual discussion

The results presented in the last section, and in particular the properties of P_{eq} , suggest that this classical model of the storage effect, with pure environmental noise, is incomplete and leads inevitably to a conceptual breakdown. In this section these difficulties are presented, and the inclusion of demographic stochasticity into the model is suggested as a possible (and plausible) solution.

As seen above, P_{eq} depends only on δ , the correlation time of environmental variations, and not on their strength σ_E^2 (as long as higher orders in σ_E^2 may be neglected). However, the environment is fluctuating on every timescale: the wind changes its velocity and direction, clouds cover the sky and cast a shadow, temperature changes slightly and so on. All these processes have correlation times of minutes or hours but their effect is minute. Still, under the rules that govern the lottery game, such processes have the ability to induce stability in the long run since $\delta \rightarrow 0$ and the amplitude of fitness fluctuations is irrelevant.

Clearly, a reasonable model should yield a P_{eq} that depends on σ_E^2 . It is highly implausible that infinitesimal changes in wind direction or in temperature play an essential role in stabilizing natural communities, no matter how much time the system is allowed to relax. A lower cutoff below which environmental variations are negligible has to be introduced. For example, one may suggest that the minimal correlation time that the model has to take into account is the time between two consecutive deaths of individuals, since all the events that affect the fitness of species between two deaths are integrated to determine the success probability of every seedling competing to replace a dead tree. By doing that one already introduces the discreteness of individuals into the model. Demographic stochasticity, which is the endogenous noise associated with the discreteness of the birth-death process, provides the natural mathematical tool to deal with these aspects of reality. Quantifying the strength of demographic noise by the parameter σ_D^2 (the value of this parameter is discussed below), we expect that the equilibrium distribution (4) should be obtained, from a general theory with demographic fluctuations, in the limit $\sigma_D^2/\sigma_E^2 \rightarrow 0$, but for any finite demographic

noise the ratio between σ_E^2 and σ_D^2 should enter the expression for P_{eq} .

Another aspect of the result (4), which also provides a hint about the importance of discreteness, is the transition between the single peak, balanced distribution at small values of δ and the imbalanced distribution at large δ . Mathematically speaking, as long as the distribution is normalizable the solution is legitimate, so the theory holds for all $\delta < 1$ and breaks down only when $\delta = 1$, where P_{eq} diverges like x^{-1} at the edges. However in practice, when the number of individuals has to be an integer, this formal approach may be misleading. If the overall size of the community is J individuals, the case $x < 1/J$ should be considered as extinction. No matter what δ is, in the long run one of the two species inevitably goes extinct and the system reaches fixation. This feature is missing in the lottery game, where all positive values are allowed for x . Most importantly, when $\delta > 1/2$, the time to fixation is short and there is no timescale over which demographic fluctuations can be ignored.

Accordingly, the stability of the system depends not only on the shape of P_{eq} , but also on the rate at which the abundance of a single species scans through all the values of x and reaches values below $1/J$, a feature that depends strongly on σ_E^2 (Kessler et al., 2015). Even if $P_{eq}(x)$ is very small for x close to zero in the regime $\delta < 1/2$, under strong environmental noise the species' abundance samples the whole phase space on relatively short timescales, leading to fast fixation. As we shall see, since the decay of P_{eq} at the edges is a power law at best, one cannot neglect extinctions even when J is large.

Demographic stochasticity has two aspects. First, it opens the possibility of extinction by allowing a species to reach an absorbing state at zero concentration. Second, it provides another source of noise, which scales like the square root of the population size, as opposed to the linear scaling that characterizes environmental stochasticity (Kalyuzhny et al., 2014b; Lande et al., 2003). These two aspects of demographic noise are of importance to the study of the storage effect, and they manifest themselves in TNTB. However, before considering TNTB we would like to address the possible relevance of the storage effect, in its traditional form with only environmental stochasticity, to the statistics of high-diversity assemblages.

4. The lottery game for many species

The applicability of the storage effect as a possible explanation for an empirical system with tens and hundreds of species was considered by Hubbell (2001), during the introduction of the neutral model. The observed species abundance distribution in the tropical forest is very wide, with support over a few decades of abundance; Hubbell argued that the prediction for a system stabilized by the storage effect is a narrow species abundance distribution (SAD) with a Gaussian-like peak around some typical value. Accordingly, Hubbell concluded that the storage effect is inappropriate for explaining patterns of species diversity in the tropical forest. Since most of the diverse communities are characterized by a hollow curve of species abundances (Ulrich et al., 2010) with many rare species and a few common ones, this argument suggests that the storage effect plays at best only a minor role in their dynamics.

In this section we will show that, in the limit of weak environmental stochasticity considered above, when the number of species S is much larger than one, the storage effect yields a Gamma-like distribution for the SAD. The Gamma distribution is known to be mathematically flexible, it fits many empirical SADs and indeed it may resemble very closely the zero-sum multinomial distribution proposed by Hubbell. Furthermore, it contains

the commonly observed Fisher log series SAD as a limiting case (see e.g., [Azaele et al., 2006, 2015](#); [Connolly et al., 2014](#)).

The generic, S species, generalization of Eq. (1) is,

$$x_i^{t+1} = (1 - \delta)x_i^t + \delta \frac{f_i x_i^t}{\sum_{j=1}^S f_j x_j^t}. \quad (5)$$

The solution for P_{eq} was obtained by [Hatfield and Chesson \(1997\)](#) (see also [Gillespie, 1980](#)),

$$P_{eq}(x_1, x_2, \dots, x_S) = (x_1 x_2 \dots x_S)^{\frac{2}{S}(\frac{1}{\delta}-1)-1}. \quad (6)$$

This is a very strong result, although it has not received its due attention in the literature. To compare directly the expression (6) to observed SADs we have extracted the single species probability distribution function by integrating out $S - 1$ species to obtain, say, $P_{eq}(x_i)$ (since all species are symmetric, we will denote it by $P_{eq}(x)$). The result is,

$$P_{eq}(x) = A x^{\frac{2}{S}(\frac{1}{\delta}-1)-1} (1-x)^{\frac{2(S-1)}{S}(\frac{1}{\delta}-1)-1}, \quad (7)$$

where A is a normalization factor.

Eq. (7) is an exact formula that reduces to (4) where $S=2$, but in this section we are interested in its implication for $S \gg 1$. In this parameter regime we need consider only $x \ll 1$, since the typical fraction of a single species never substantially exceeds $1/S$, as we shall see.

As mentioned, when $S \gg 1$, P_{eq} takes a Gamma distribution (power law followed by an exponential cutoff) form,

$$P_{eq} \approx A x^{\alpha-1} e^{-\beta x}, \quad (8)$$

where the rate factor β appears as the $x \ll 1$ limit of

$$(1-x)^{\frac{2(S-1)}{S}(\frac{1}{\delta}-1)-1} \approx e^{-\beta x}, \quad (9)$$

with $\beta = (2/\delta) - 3$. The shape factor $\alpha = \frac{2}{S}(\frac{1}{\delta} - 1)$ will be greater than one (meaning that the distribution vanishes at zero and has a peak in the vicinity of $1/S$) if

$$\delta < \delta_c = \frac{2}{S+2}, \quad (10)$$

while for $\delta > \delta_c$ the distribution diverges at zero but is still integrable. These behaviors are illustrated in [Fig. 2](#). When α is small the distribution approaches the Fisher log series, and in general it corresponds to the generalized Fisher log series distribution that was discussed in [Kessler and Shnerb \(2014\)](#). Note that if $\delta > 2/3$ the assumption $x \ll 1$ does not hold anymore; we will not consider this case here.

For a fully surveyed empirical community the species richness S is given, so the only parameter in (7) which is left to be fitted is δ . This makes the fit less impressive, of course, but the model is more parsimonious and its results may be preferred, e.g., when applying Akaike information criterion that includes a penalty to discourage overfitting, over two parameter theories.

In particular, to explain the observed SAD on the BCI plot as reflecting a community that acquires its stability from the storage effect, we first look at [Fig. 3b](#) and find the range of abundances for which the decay of the SAD is exponential, this indicates what $1/\beta$ is. In the BCI (dbh > 10 cm) this “knee” (the range of abundances in which the SAD, on a double logarithmic scale as in [3b](#), curves down) is definitely below 500 trees, i.e., below 2% of the forest population (around 21,000 trees). Since this scale is determined by $1/\beta$ where $\beta = (2/\delta) - 3$, the correlation time δ needed to explain the SAD in the Barro-Colorado plot has to be below a fraction 0.03 of a generation. This estimate works quite nicely: [Fig. 3](#) shows the observed SAD for BCI trees together with the P_{eq} presented in Eq.

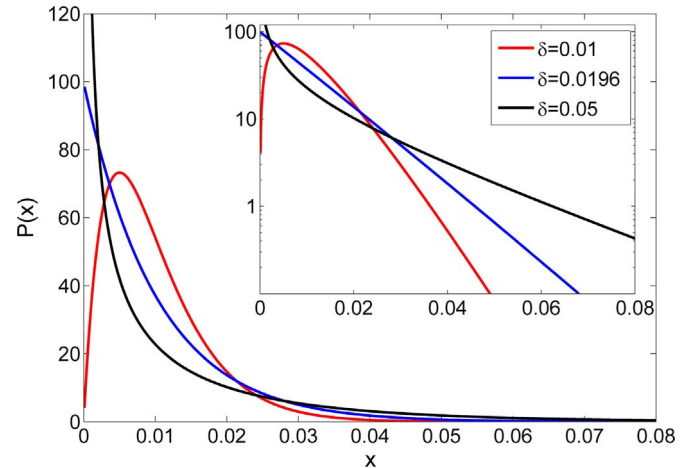


Fig. 2. The species abundance distribution $P_{eq}(x)$, given by Eq. (7) for $S=100$ species, is plotted here vs. x for three values of δ . In all cases the SAD approximates a Gamma distribution, with shape factor $\alpha > 1$ for $\delta < \delta_c = 0.0196$ and $\alpha < 1$ if $\delta > \delta_c \approx 0.0196$. At the critical δ the power vanishes and the distribution has an almost pure exponential decay as demonstrated in the semi-logarithmic plot in the inset.

(7), where the only fitted parameter is δ . Indeed the best fit was obtained for $\delta = 0.027$, as expected. One can see that this one parameter fit is worse than the two-parameter fit using zero-sum multinomials (ZSM) statistics, but it is not unacceptable. Interestingly, when plotted using the double-logarithmic (Pueyo) plot instead of Preston plot the single parameter fit using (7) looks much better.

However, the value $1/40$ (of a generation) for δ appears to be unrealistic. As mentioned in [Kalyuzhny et al. \(2015\)](#) the value of δ has been found to be around $1/5$, and the order of magnitude in difference is too large to be neglected. Similarly, estimations of the corresponding numbers for trees in the whole Amazon basin ([Ter Steege et al., 2013](#)) suggest that the most abundant species constitutes about 1% of the population (meaning that the “knee” appears at even smaller relative abundances) and the corresponding value of δ , smaller than $1/200$, again seems unrealistic.

Altogether, it appears that a storage model may provide the type of hollow-curve SADs that characterize empirical systems. Its flexibility is limited since α and β are both determined by δ , and other theories provide better fits, still one may believe that adding another parameter (in a theory that takes into account spatial effects, for example) may solve this difficulty. The need to extend the lottery model and to include demographic noise is *not* the inability of (7) to support fat-tailed SADs, but the following three arguments:

1. Conceptually, as mentioned above, we would like to find an SAD that depends on σ_E , not only on δ .
2. The empirical SAD (and the theoretical expression (7) with the δ values that yield a decent fit) has support on small absolute numbers, meaning that extinction events must occur and are important, or, equivalently, that demographic noise must be taken into account.
3. The correlation time of the environmental variations needed to account for empirical datasets appears to be unrealistically short.

In any case, once extinctions are incorporated into the model one should include speciation events to balance the species richness; the resulting model is the TNTB which is discussed in the next section.

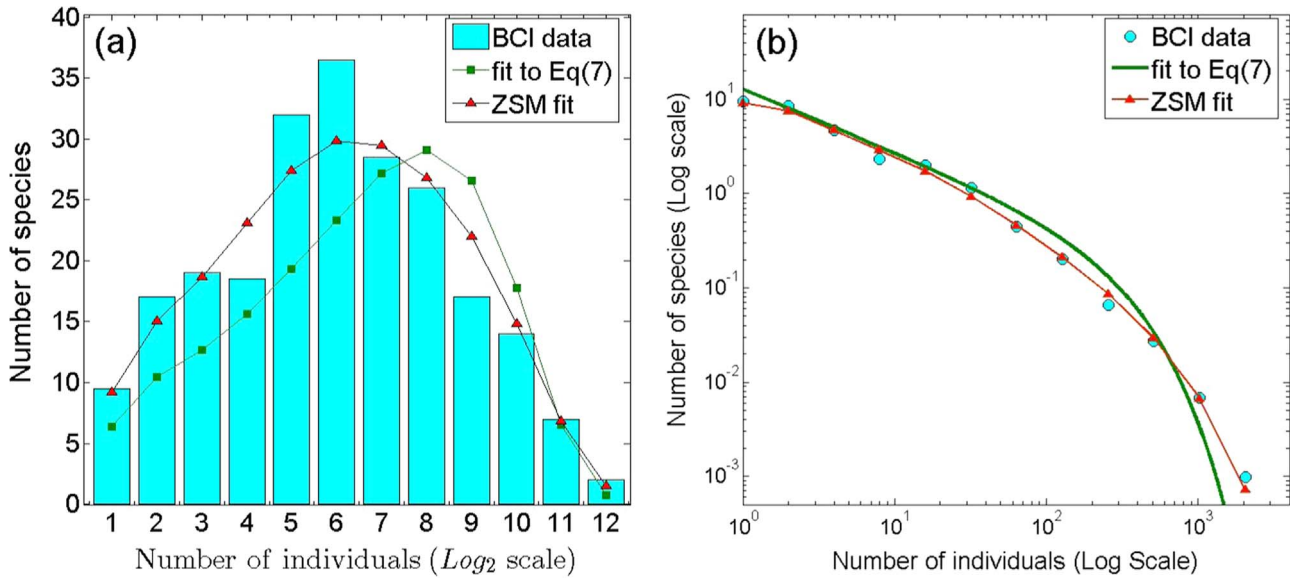


Fig. 3. Explaining the BCI plot species abundance distribution. The species abundance distribution of the BCI (1995 census) tropical forest is presented (light blue bars in the Preston plot (a), light blue circles in the double logarithmic (Pueyo) plot (b)), together with the best fit to the SAD predicted for a lottery game with $\delta = 0.0265$ (Eq. (7)) and the ZSM theory (parameters taken from Volkov et al. (2003), $\theta = 47.22$, $m = 0.1$). Although the ZSM appears to give a better fit, it is clear that the Chesson–Warner lottery game does provide the kind of hollow curve SADs that appear in empirical studies. Moreover, for the ZSM two parameters were used, whereas only one parameter was employed for the lottery game. (For interpretation of the references to color in this figure caption, the reader is referred to the web version of this paper.)

5. The TNTB: storage effect, demographic noise and speciation

The time-averaged neutral theory of biodiversity deals with a community of species, all having the same average fitness. Species are subject to demographic and environmental stochasticity. Under demographic noise species may go extinct, and these extinction events are balanced by speciation. Four parameters govern the results: in addition to δ and σ_E , the correlation time and the amplitude of environmental variations, here one should take into account the per-birth chance of speciation ν , and the strength of the demographic noise, σ_D . One should introduce these two processes together: without extinction, speciation will cause the number of species to grow unboundedly. Without speciation, demographic noise will lead to fixation by a single species in the long run.

The standard way to introduce speciation is to assume that an offspring carries the taxonomic identity of its mother with probability $1 - \nu$, and is the originator of a new species with probability ν (see details of our simulation technique in Appendix B). The demographic stochasticity is characterized by σ_D^2 , the variance of the number of offspring per individual and usually takes a value between 2 (for a geometric distribution of offspring) and 1 (for a Poisson distribution). In the limit $\sigma_E = 0$, without environmental noise and storage effect, one obtains the metacommunity version of Hubbell's neutral theory (or Kimura's neutral model), where P_{eq} (for a high-diversity system with $1/J < x \ll 1$) is given by Fisher log-series,

$$P_{eq}^{\sigma_E=0}(x) = \frac{A}{x} e^{-2\nu x / \sigma_D^2}, \quad (11)$$

where A is a normalization constant. The species richness S reflects the balance between extinction and speciation

$$S^{\sigma_E=0} = -\frac{2\nu}{\sigma_D^2} \log\left(\frac{2\nu}{\sigma_D^2}\right) J. \quad (12)$$

What happens in TNTB, when the storage mechanism acts together with demographic noise and speciation events? Here we would like to emphasize a few generic features of this system:

1. Demographic stochasticity allows for extinction while speciation increases species richness, and the balance between these two processes is determined by δ and σ_E . The lower the value of δ , the sharper the SAD peak in the vicinity of $1/S$, the time to extinction of a single species increases, and the chance of a low-abundance species to invade grows. Accordingly, the steady state species richness S decreases monotonically with increasing δ . Dynamically, when the chance of extinction is low the process of speciation acts to increase the number of species S , δ_c decreases (see Eq. (10)) and the probability mass of P_{eq} in the region $x \ll 1$ grows, leading to an increased rate of extinction until it balances the effect of speciation and the system reaches a steady state at finite S .
2. The inclusion of both demographic and environmental noise introduces a new scale into the problem. As discussed in Kessler and Shnerb (2014) and Kessler et al. (2015), as long as $x \ll \sigma_D^2 / \sigma_E^2$ the dynamics of a species is dominated by demographic noise, while, above this value, environmental variations are more important. Accordingly, as one can see in Fig. 4, P_{eq} of the TNTB has two regimes. For large x one observes (8), the storage power-law $\alpha - 1$ followed by an exponential cutoff. For $x < \sigma_D^2 / \sigma_E^2$ the power $\alpha - 1$ is replaced by a $1/x$ dependence, a characteristic of the Fisher log-series. Therefore, the conceptual problem raised in Section 3, namely the fact that the theory of the storage effect predicts P_{eq} to be independent of σ_E , is solved within the TNTB framework: the ratio σ_D^2 / σ_E^2 determines the crossover from the $1/x$ decay to the behavior described by (8), and the SAD (and the overall species richness S) does depend on σ_E . In the $\sigma_E \rightarrow 0$ limit the TNTB converges to the standard neutral theory of Hubbell.
3. Given that, one may wonder about the effect of environmental stochasticity on species richness. On the one hand, σ_E is responsible for the storage effect that provides stability and allows for low abundance species to invade. On the other hand (see Eq. (14) below and the following discussion), in systems without a storage effect (Kessler and Shnerb, 2014; Kessler et al., 2015), environmental stochasticity clearly acts to lower S , as it increases the rate of extinction events since environmental

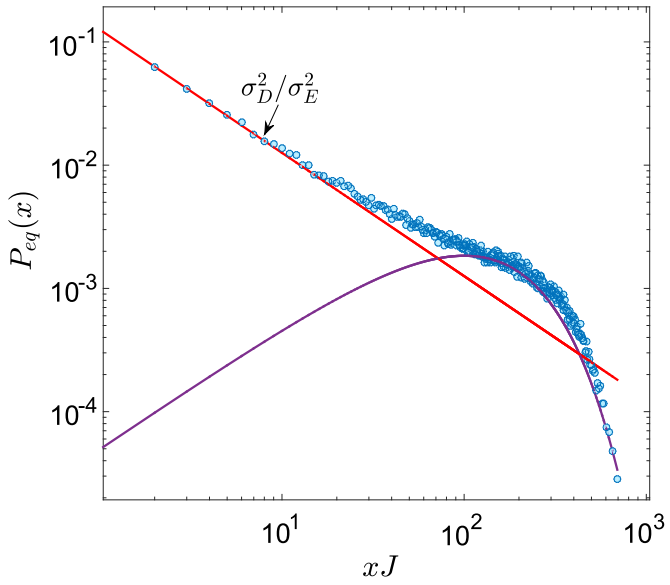


Fig. 4. The light blue circles represent the SAD obtained from a simulation of the TNTB model (see Appendix B), with $\sigma_D^2 = 2$, $\sigma_E^2 = 0.25$, $\delta = 10^{-4}$, $\nu = 10^{-3}$ and $J = 10^4$. The results represent an average over time, abundances have been recorded every 10^6 elementary timesteps. The red straight line has a slope -1 , and it fits the data perfectly up to σ_D^2/σ_E^2 , the region dominated by demographic stochasticity. The Gamma distribution, with $\alpha = 2$ and $\beta = 0.01$ is shown by the purple curve. The parameters α and β were obtained from the fit to the data, and differ from the predictions of the storage theory, since the demographic noise and mutations significantly weaken the storage effect. (For interpretation of the references to color in this figure caption, the reader is referred to the web version of this paper.)

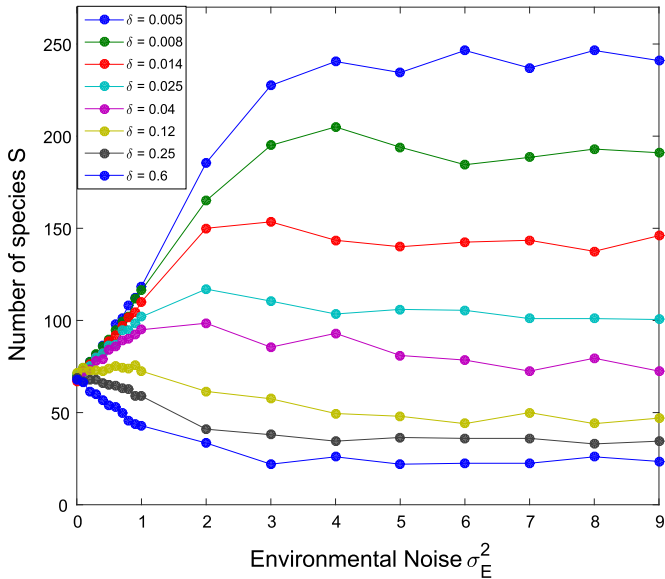


Fig. 5. S , the species richness, is plotted against σ_E^2 , the amplitude of the environmental stochasticity. The results were obtained in simulations of a TNTB community with $J = 10^4$, $\sigma_D^2 = 2$, $\nu = 10^{-3}$ for various environmental correlation times δ (given in the legend in units of one generation). S reflects the balance between extinction and speciation; the lower is δ , the stronger is the storage effect and thus S increases. An increase in the strength of environmental variation σ_E^2 may either decrease S (since it increases abundance variation) or increase the species richness by facilitating the storage effect. Here we see that the general trend depends on the value of δ . All the lines converge to the NTB limit, $S = -J\nu\log(\nu) \cong 70$, when $\sigma_E \rightarrow 0$.

fluctuations cause a species to visit more frequently the dangerous zone of low abundance.

Fig. 5 solves this puzzle: it shows that the effect of environmental stochasticity on species richness, when all other

parameters are kept fixed, is determined by the correlation time δ . For small δ 's the storage effect wins and in general the species richness increases with the amplitude of environmental variations. For large values of δ the increase in the system's variability leads to a decrease in S .

4. The NTB was criticized by many authors for its strict commitment to perfect neutrality (Zhou and Zhang, 2008). Under the rules of the neutral game, even the slightest fitness difference leads to a fixation of the system by the fittest species (in the absence of speciation) or to the appearance of an SAD that reflect Darwinian dominance, with one common species that occupies most of the community and a few rare, short lived, species (Kessler and Shnerb, 2014). The stabilizing effect of the storage mechanism resolves this difficulty. Even if the average fitness of different species is *not* the same, the system may still support high diversity.

To demonstrate this we have simulated the non-neutral modification of the TNTB, when the expression for fitness, Eq. (2), is replaced by

$$f_i = e^{\eta_i + \gamma_i^t}, \quad (13)$$

where η_i is a time independent, species specific component of the fitness of the i -th species, taken from a Gaussian distribution with standard deviation Q and zero mean. η_i reflects the mean tendency of the environment to favor, or disfavor, species i . When $\sigma_E \rightarrow 0$ the introduction of these time independent fitness differences leads to a biodiversity collapse, as seen in Fig. 6. However, as σ_E increases, the number of species grows since the storage effect induces stability. Finally at large σ_E the effect of fitness differences disappears and the species richness takes its $Q=0$ value, as in the TNTB.

An interesting feature of the finite Q dynamics is the unimodal dependence of S on σ_E^2 when δ is sufficiently large. While weak environmental stochasticity stabilizes the species, strong variations lead to faster extinction and reduce S and biodiversity reaches a maximum under *intermediate disturbance*. One may expect such an effect in nonadditive systems (see Fox, 2013),

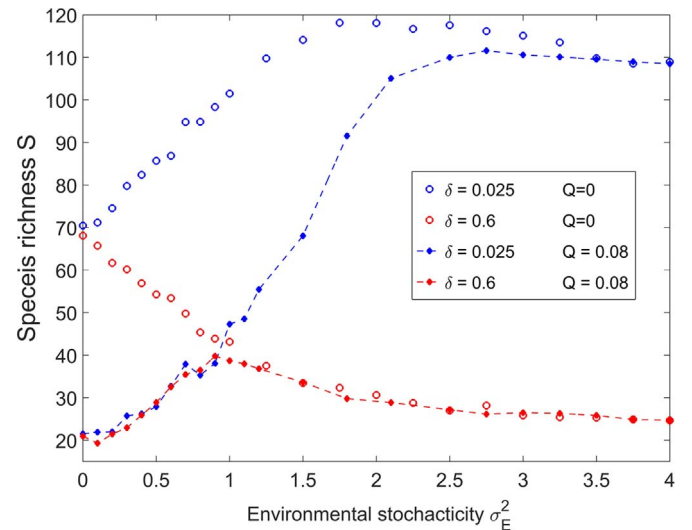


Fig. 6. S , the species richness, is plotted against σ_E^2 , amplitude of the environmental stochasticity for simulations with $Q=0$ (same as in Fig. 5, empty circles) and for the case with time independent fitness differences (η_i s) with $Q=0.08$ (filled circles connected by dashed lines). All other parameters are the same used in Fig. 5. When $\sigma_E^2 = 0$ time independent fitness differences lead to a biodiversity collapse, where the fittest species is dominant and all other species are rare. On the other hand, when σ_E^2 is large environmental noise washes out the effect of time independent fitness differences and $S(Q)$ approaches $S(Q=0)$. The species richness peaks at intermediate level of disturbance for $\delta = 0.6$.

and we believe that our model provides an appropriate framework for its analysis.

Before concluding this section, we would like to stress that environmental stochasticity and the storage effect are not synonymous. When the environmental variations affect only the death rate, or when $\delta = 1$, there is no storage effect and σ_E is a purely destabilizing factor. Even a slight modification of the rules governing the process may kill the storage effect. For example, in the lottery game all the trees in the forest are competing for an open gap, where the chance of a species to win depends on its fitness. The pairwise competition version of the same game, where two individuals are picked at random and an offspring of one replaces the other with probability that depends on their relative fitness, has no significant storage effect. With respect to this specific duel all other trees play no role, so it corresponds to the $\delta = 1$ case of the lottery game. Conversely, in the NTB limit (without environmental stochasticity) there is no difference between these two versions of the neutral game, see e.g. Volkov et al. (2003).

The SAD for TNTB *without* the storage effect (e.g., for the pairwise competition case) for $1/J < x \ll 1$ was calculated in Kessler and Shnerb (2014),

$$p_{eq}^{no\ storage}(x) = \frac{C}{x} \left(1 + \frac{2\sigma_E^2}{\sigma_D^2} x \right)^{-1-\nu/\sigma_E^2} \quad (14)$$

where C is a normalization constant. Here the power law decay $1/x$ that characterizes the region dominated by demographic noise is replaced, for $x > \sigma_D^2/\sigma_E^2$, by a power law with a larger exponent $2 + \nu/\sigma_E^2$.

6. Summary and discussion

Mechanisms that maintain species diversity are usually classified according to their stability properties. Some mechanisms provide a stable equilibrium, while in other mechanisms the dynamics of each species is unstable and the diversity reflects a balance between extinction and speciation/immigration. This distinction is related to timescales: under the inevitable influence of demographic noise every species eventually goes extinct, however in models that support a stable equilibrium the extinction time is exponential in the species' abundance, while under unstable equilibrium, as in the neutral model, the time to extinction scales linearly with the abundance. To maintain the diversity of a metacommunity these timescales should be comparable with the evolutionary timescale that determines the rate at which new species enter the system and balance the diversity losses due to extinction.

A system that acquires its stability due to the storage effect is somewhere in-between. The stabilization is based on environmental stochasticity, which is, at the same time, a destabilizing force. As we have seen, the outcome of the competition between these two aspects of the same phenomenon—environmental stochasticity—is determined by one parameter, δ , the correlation time of the environment. If δ is large the destabilizing effects dominate and environmental stochasticity reduces biodiversity. When δ is small, as seen in Fig. 5, the stabilizing effect associated with the storage mechanism leads to an increase of extinction times and the overall biodiversity.

In the metacommunity version of Hubbell's neutral theory, speciation and demographic drift are the only factors that govern the dynamics of the community, leading to the Fisher log-series SAD and species richness which is given by (12). The $1/x$ decrease of the SAD at small x does not fit the observed statistics on, say, the

Barro-Colorado Island and other local communities, where the slope is clearly weaker than $1/x$ (in a Preston plot, where the number of species in any abundance octave is plotted without normalization by the width of the octave, $1/x$ is translated into a straight horizontal line, while the Preston plots of empirical local communities show a unimodal behavior, see Fig. 3a). To account for that, in the mainland-island version of NTB the statistics of a local community are governed by two parameters, the fundamental biodiversity number of the metacommunity $\theta = 2\nu_M/\sigma_D^2$ and the chance of migration to the island. The emerging zero-sum multinomial SAD fits the empirical evidence, as may be seen in Fig. 3 above. Nevertheless the dynamics, in particular the rate of abundance variations, is too fast to be explained by the neutral model (Kalyuzhny et al., 2014a, b, 2015).

TNTB, which was shown to explain both static and dynamic patterns (Kalyuzhny et al., 2015), has three extreme limits. When $\sigma_E \rightarrow 0$ it converges to the NTB, as environmental variations vanish. When $\sigma_D/\sigma_E \rightarrow 0$ it converges to the classical lottery model of Chesson and Warner. The other limit is $\delta \rightarrow 1$, when environmental noise does affect the system but there is no storage. The SADs in these three limits were presented in this paper (Eqs. (7), (11), (14)).

In between, as showed in Section 4, the situation is more complicated, and the way environmental stochasticity affects species richness is determined by the correlation time δ . For short correlation times, S is an increasing function of σ_E , while for longer correlation times the situation is closer to the one discussed in Kessler and Shnerb (2014) – a species may enjoy a long time in which its population grows, so the SAD widens and the overall species richness S decreases when environmental variations increase in amplitude.

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Appendix A. Derivation of the Fokker–Planck equation (3)

Our starting point is the recursion relation, Eq. (1).

$$x_1^{t+1} = x_1^t(1 - \delta) + \delta \frac{f_1^t x_1^t}{f_2^t x_2^t + f_1^t x_1^t} \quad (A.1)$$

To translate this equation to the Fokker–Planck language, one would like to find the average change in x_1 between t and $t + 1$, $\overline{\Delta x_1}$, and also its variance $Var(x_1) = \overline{(\Delta x_1)^2} - (\overline{\Delta x_1})^2$. The Fokker–Planck equation then takes the general form,

$$\frac{\partial P(x_1, t)}{\partial t} = - \frac{\partial}{\partial x_1} \left(\overline{\Delta x_1} P(x_1, t) \right) + \frac{1}{2} \frac{\partial^2}{\partial x_1^2} \left(Var(x_1) P(x_1, t) \right) \quad (A.2)$$

The transition between (A.1) and (A.2) is justified if both the mean change and its variance are small (in comparison with x_1) during a single timestep, such that a differential equation is an appropriate description of the process, see below.

To calculate $\overline{\Delta x_1}$, Eq. (A.1) should be written as (we have replaced x_2 by $1 - x_1$, so from now on we regard x_1 as x).

$$\Delta x = x^{t+1} - x^t = x^t \delta \left(\frac{1}{(f_2/f_1)(1 - x^t) + x^t} - 1 \right) \quad (A.3)$$

Writing $f_i = e^{\gamma_i}$, and expanding Δx in powers of $\Delta \gamma \equiv \gamma_1 - \gamma_2$, we have

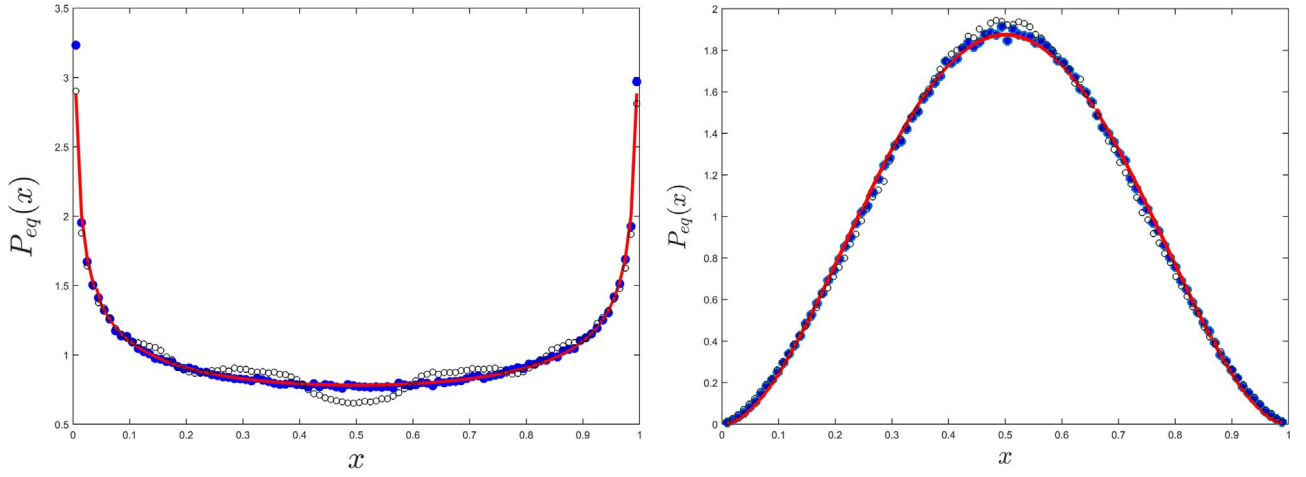


Fig. A1. The predictions of the Fokker–Planck Eq. (4) for $P_{eq}(x)$ (full line), are compared with P_{eq} obtained from numerical simulations of the recursion relation (1) for different values of δ and $\delta\sigma_E^2$. For $\delta = 0.6$ (left), the simulated P_{eq} at $\sigma_E^2 = 1$ (filled blue circles) is almost indistinguishable from the Fokker–Planck predictions, while for $\sigma_E^2 = 6$ (open black circles) one observes a few little wiggles. For $\delta = 0.25$ (right panel) the situation is even better, at $\sigma_E^2 = 1$ the approximation is exact and only for $\sigma_E^2 = 100$ (open black circles) slight deviations may be identified. For smaller values of δ we were unable to identify any deviations even for larger values, up to $\sigma_E^2 = 10000$. (For interpretation of the references to color in this figure caption, the reader is referred to the web version of this paper.)

$$\Delta x = x\delta \left[\Delta\gamma(1-x) + \frac{1}{2}(1-x)(1-2x)(\Delta\gamma)^2 + O((\Delta\gamma)^3) \right], \quad (\text{A.4})$$

where the negligence of higher orders in $\Delta\gamma$ is justified if this quantity is small compared to one. Since $\overline{\Delta\gamma} = 0$ the first term on the r.h.s. of (A.4) vanishes. The strength of the environmental stochasticity is defined via $\sigma_E^2 \equiv (\Delta\gamma)^2$, so we are left with,

$$\overline{\Delta x} \approx \frac{\delta\sigma_E^2}{2}x(1-x)(1-2x). \quad (\text{A.5})$$

Writing $(\Delta x)^2$ as the square of the r.h.s. of (A.3), expanding to second order in $\Delta\gamma$, averaging over $\Delta\gamma$ and collecting the non-vanishing quantities one obtains,

$$\overline{\Delta x^2} \approx \sigma_E^2\delta^2x_1^2(1-x_1)^2 \quad (\text{A.6})$$

Finally, keeping only second order terms in $\Delta\gamma$, $\text{Var}(\Delta x) \approx \overline{\Delta x^2}$, and by plugging the results of (A.5) and (A.6) into (A.2) one obtains Eq. (3) of the main text.

The replacement of $P(x_{t+1}) - P(x_t)$ by the time derivative of P is allowed if the changes in x during one timestep are small, i.e., if $\delta\sigma_E^2 \ll 1$. However, when comparing the results obtained from direct simulations of the recursion relation (1) and the predictions of the Fokker–Planck equation, say, (4), one finds that this approximation holds in a much wider regime, see Figs. A1.

Note that, as long as the transition to the Fokker–Planck equation (replacement of the recursion relation by a continuous in time differential equation) may be justified, the limit $\sigma_E \rightarrow \infty$ is quite trivial. In this limit with probability $1/2$ $x_{t+1} = x_t(1-\delta)$ and with probability $1/2$ $x_{t+1} = x_t + \delta(1-x_t)$. Accordingly, $\overline{\Delta x} = \delta(1-2x)/2$ and $\text{Var}(x) = \delta^2/4$. The resulting Fokker–Planck equation has an x -independent diffusion term: it does not vanish on the edges, since growth is independent of the abundance. This leads to

$$P_{eq}(x) \sim e^{\frac{4x(1-x)}{\delta}}, \quad (\text{A.7})$$

meaning that the distribution in this limit is a Gaussian around $x = 1/2$. In fact, writing $y = x - 1/2$ and expanding for small y both (4) and (A.7) converge to $\exp(-4y^2/\delta)$, and for small δ (where the deviations from $x = 1/2$ are small) higher orders in y are negligibly small. Therefore (see Fig. A1 and its caption) as long as δ is small it is quite difficult to find any difference between the prediction of (4) and the results of simulations at large σ_E^2 .

A similar argument for S species yields $P_{eq}(x) \sim \exp(-(x - 1/S)^2/w)$ where $w = S/(\delta[1 - 1/S])$.

Appendix B. Simulating the dynamics of a community

Our individual-based simulation technique is a simple generalization of the one used in simulations of the neutral model for a well-mixed (panmictic) community.

We start with J individuals, each belonging to one of S species, and each species i has a fitness $f_i = \exp(\gamma_i)$, where the values of γ_i are picked from a normal distribution with zero mean and width σ_E .

During an elementary timestep one individual is chosen (at random, independent of its species affiliation) to die. If the dying individual is a singleton, the number of species is reduced by one. Then, with probability ν , this individual is replaced by the originator of a new species, the number of species in the system grows by one and a fitness is assigned to the new species using the procedure described above.

With probability $1 - \nu$ the individual is replaced by a descendant of another individual that inherits its parent's identity. The chance of the species i to capture the empty slot is given by $f_i x_i / \sum_{j=1}^S f_j x_j$, where x_i is the abundance of species i (an integer). After each elementary timestep the time is incremented by $1/J$.

When the time reaches δ (after δJ elementary steps) all the fitness values f_i are picked for all species using the same procedure, with no correlation between the new and the old fitness. This implies that the correlation time of the environmental variations is a fraction δ of a generation.

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