

# Simultaneous adsorption of $\text{Hg}^{2+}$ , $\text{Cd}^{2+}$ and $\text{Cu}^{2+}$ ions from aqueous solution with mesoporous silica/DZ and conditions optimise with experimental design: kinetic and isothermal studies

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This study used dithizon (DZ) and mesoporous as a sorbent for simultaneous removal of heavy metal (mercury, cadmium and copper) ions from aqueous solution and real samples. Santa Barbara Amorphous synthesised and characterised by X-ray diffraction, Fourier-transform infrared spectroscopy and Brunauer–Emmett–Teller. Heavy metal ions measured by atomic absorption spectrometry and inductively coupled plasma optical emission spectrometry. Optimised conditions such as pH, removal time, sorbent dosage and concentration analyte with central composite design and obtained 5, 20 min, 20 mg and 25 ppm, respectively. The equilibrium data were fitted to isotherm models such as Langmuir, Freundlich and Tempkin and the results revealed the suitability of the Langmuir model. The maximum adsorption capacities for  $\text{Hg}^{2+}$ ,  $\text{Cd}^{2+}$  and  $\text{Cu}^{2+}$  ions achieved 25.04, 30.3 and 31.79 mg/g, respectively. Kinetics data fitting to pseudo-first order, pseudo-second-order and Elovich models confirmed the applicability of pseudo-second-order kinetic model for the description of the mechanism and adsorption rate. This sorbent could regenerate easily with ethylenediaminetetraacetate solution.

**1. Introduction:** Heavy metal ions are one of the world current problems; since, they are toxic pollutants, and have density more than  $5 \text{ g/cm}^3$  and are widely applied in batteries, plastics, pigments, metal plating and petroleum [1–4]. The lawful amount of heavy metal ions such as  $\text{Hg}^{2+}$ ,  $\text{Cd}^{2+}$  and  $\text{Cu}^{2+}$  in water according to World Health Organisation around 0.001, 0.003 and 2 mg/l, respectively [5].

High level of heavy metal ions especially cadmium, copper, mercury, silver and radioactive ions in aqueous solution are harmful to humans and ecosystems [6]. Such as mercury ion could pass skin, attracted central nervous system and effect on chromosomes [7, 8]. Cadmium ions could damage bones, lungs, livers, causing diarrhoea and metabolic disorders [9, 10]. Copper ions effect on eyes, bones, stomach and caused Menkes syndrome and Alzheimer's disease [11, 12]. Thus, we need that remove these pollutants. The several methods exist for removal such as reduction, coagulation, ion exchange, filtration and photocatalytic oxidation, electrodeposition, electrochemical, precipitation and membrane separation that among them adsorption is the best, rapid, easy and economical method [2, 13–15].

Several materials such as mesoporous silica, activated carbon, polymers, clays, talk, zeolite, fly ash and biomass are used for adsorption [16, 17]. Porous materials classified into three kinds by their pore size: macroporous ( $>50 \text{ nm}$ ), mesoporous ( $2\text{--}50 \text{ nm}$ ) and microporous ( $<2 \text{ nm}$ ). Mesoporous silica such as Santa Barbara Amorphous (SBA-15) with unique properties such as high specific surface area, good pore diameter ( $2\text{--}50 \text{ nm}$ ), highly nanoporous structures and thermally suitable for adsorption process [18–21].

Immobilisation of dithizon (DZ) in SBA-15 has been used to improve adsorption capacity and increase active sites because DZ as the ligand with  $-\text{SH}$  (thiol) and  $-\text{NH}$  (amin) groups was coupled with metal ions [17]. In this Letter, we simultaneously remove several ions from aqueous solution with SBA-15/DZ and optimised experimental conditions by central composite design (CCD). The adsorption isotherms and kinetics have also been investigated to display the adsorption mechanisms.

## 2. Materials and methods

**2.1. Materials:** All reagents and materials were used from Merck and Aldrich Co.  $\text{Hg}(\text{NO}_3)_2$ ,  $\text{H}_2\text{O}$ ,  $\text{Cd}(\text{NO}_3)_2$ ,  $4\text{H}_2\text{O}$  and  $\text{Cu}(\text{NO}_3)_2 \cdot 9\text{H}_2\text{O}$  salts from Merck Co. utilised for provided three stock solutions (1000 ppm). Tetraethylorthosilicate (TEOS) and hydrochloric acid (HCl) solution from Merck Co. and Pluronic (P123,  $\text{Mn}=5800$ ,  $\text{EO}_{20}\text{PO}_{70}\text{EO}_{20}$ ) from Aldrich Co. was used for synthesised SBA-15. DZ as a ligand utilised from Merck Co.

**2.2. Apparatus:** Analyte concentration was determined by inductively coupled plasma optical emission spectrometry (Shimadzu, ICPS-7000) and atomic absorption spectrometry (AAS) (Shimadzu, AA-6300). Fourier-transform infrared spectroscopy (FT-IR) (Shimadzu, 8400s), X-ray diffraction (XRD) (PANalytical,  $\times$  Pert Pro MPD), Brunauer–Emmett–Teller (BET) (Micrometrics, ASAP-2020) were used for characterisation SBA-15.

**2.3. Preparation of SBA-15:** Mesoporous silica SBA-15 were synthesised by hydrothermal method [20, 22, 23]. About 4.00 g Pluronic (P123,  $\text{Mn}=5800$ ,  $\text{EO}_{20}\text{PO}_{70}\text{EO}_{20}$ ), 90.0 ml of HCl 2.00 M and 21.00 ml deionised water were mixed at room temperature. In the next step, 6.40 g TEOS was added to the mixture and heated for 24 h at  $40^\circ\text{C}$ . Then, this blend was shredded in an autoclave at  $100^\circ\text{C}$  for 24 h. After these works, the gotten mixture was filtered and washed. The white powder was calcined at  $600^\circ\text{C}$  for 6 h.

**2.4. Adsorption process:** According to Fig. 1, first, heavy metal ions solution of  $\text{Hg}^{2+}$ ,  $\text{Cd}^{2+}$  and  $\text{Cu}^{2+}$  with SBA-15 were shaken.

Second, 1.00 ml ethanolic solution consisting of DZ was added to the mixture and was shaken for a patch that formed  $[\text{M}(\text{DZ})_2]$  complex (Fig. 2). Then, the mixture was filtered and the solution below filter was analysed by ICP and AAS.

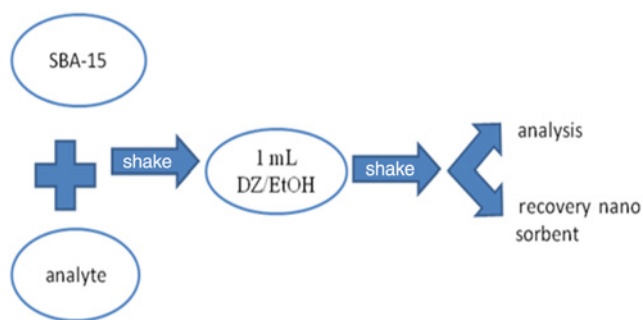


Fig. 1 Adsorption process diagram

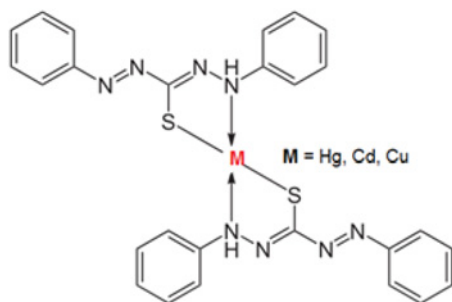


Fig. 2  $[M(DZ)_2]$  complex

Table 1 Factors and their levels for the CCD

Factor	Symbols	Levels				
		$-\alpha$	$-1$	$0$	$+1$	$+\alpha$
pH	A	4.0	4.5	5.0	5.5	6.0
dose of adsorbent, mg	B	20.00	25.00	30.00	35.00	40.00
shaking time, min	C	20.00	22.50	25.00	27.50	30.00

Finally, adsorption capacity was calculated from the equation below:

$$q = \frac{V(C_0 - C_t)}{M} \quad (1)$$

$q$  is the adsorption capacity (mg/g);  $V$  is the volume of the solution (L);  $C_0$  is the initial concentration (mg/l);  $C_t$  is the equilibrium concentration (mg/l); and  $M$  is the amount of sorbent (g).

2.5. Design of experiments: The CCD was employed to investigate the effects of different operating factors on Cd(II), Hg(II) and Cu(II) adsorption capacity and reveal the optimum conditions for Cd(II), Hg(II) and Cu(II) removal as well as build models [24, 25]. By using the CCD method, 30 experiments (including six repetitions at a central point) were designed. The factors (variables) were: pH, a dose of adsorbent, shaking time and the ions concentration. For each factor, five levels were determined. These values are shown with  $-\alpha$ ,  $-1$ ,  $0$ ,  $+1$  and  $+\alpha$  codes in Table 1 ( $\alpha=2$ ).

### 3. Result and discussion

3.1. Characterisation: XRD pattern of the mesoporous silica SBA-15 and SBA-15/DZ is shown in Fig. 3. A strong peak (1 0 0) and two weak peaks (1 1 0) and (2 0 0) indicated that SBA-15 have hexagonal structure with symmetry P6mm. Constant peaks in SBA-15/DZ indicates modification of SBA-15 not affected in influence. Absorption of DZ caused decreases of intensity.

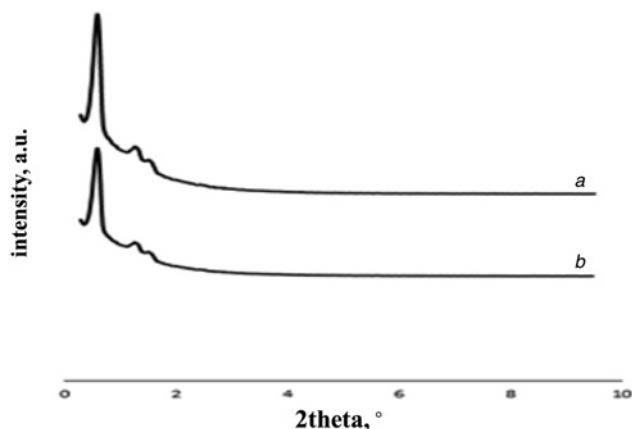


Fig. 3 XRD pattern of the  
a SBA-15  
b SBA-15/DZ

Here,  $d_{100}$  achieved from XRD pattern was 9.05 nm. Then, the lattice parameter ( $a_0$ ) was calculated and achieved 12.5 nm. The lattice parameter is estimated from the equation below:

$$a_0 = \frac{2d_{100}}{\sqrt{3}} \quad (2)$$

$a_0$  is the lattice parameter (nm) and  $d_{100}$  is the distance of page (nm).

Fig. 4 shows BET analysis and sorption/desorption  $N_2$  isotherm. BET analysis results were summarised in Table 2. Pore diameter and a wall thickness of mesoporous silica were achieved 85 Å (8.5 nm) and 4 nm, respectively. Absorption DZ on surface SBA-15 caused decreases total pore volume.

FT-IR spectrum results showed peaks in 3435, 1630, 1083 and 960  $\text{cm}^{-1}$  corresponding to stretching vibration O-H bond in silanol group, vibration  $\text{H}_2\text{O}$  molecules that adsorption, vibration

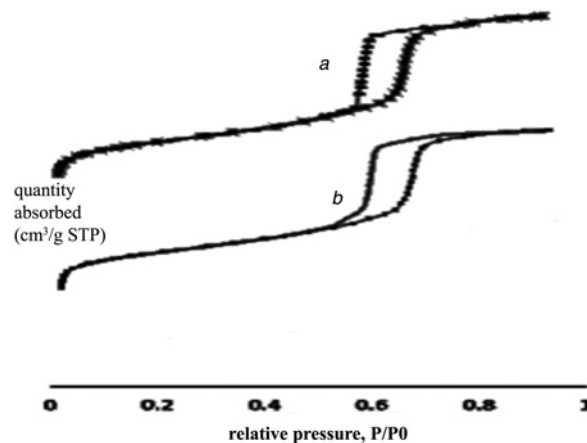


Fig. 4 Sorption/desorption  $N_2$   
a SBA-15  
b SBA-15/DZ

Table 2 BET analysis of the SBA-15 and SBA-15/DZ

Sample	Surface area, $\text{m}^2/\text{g}$	Pore diameter (Barrett, Joyner and Halenda (BJH)), Å	Total pore volume, $\text{m}^3/\text{g}$
SBA-15	869	85	1.034
SBA-15/DZ	586	76	0.96

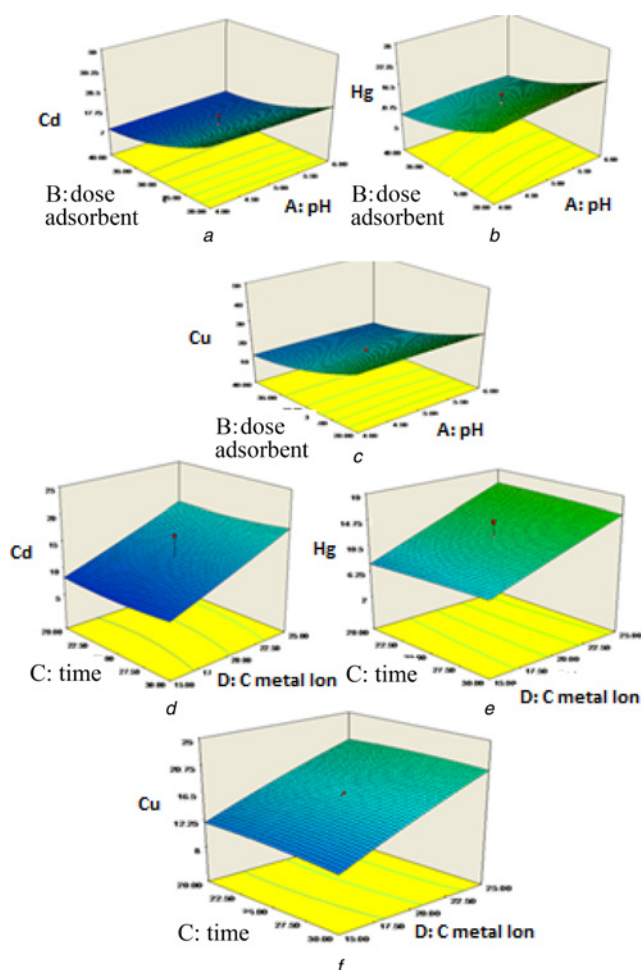
Si–O–Si and Si–O bond in silanol group, respectively. Here, 1435 and 1532  $\text{cm}^{-1}$  relative to C=S and C=N in SBA-15/DZ, respectively.

3.2. Modelling of adsorption: Experimental conditions are optimised by CCD. In the present work, the effects of four factors (pH, dosage adsorbent, removal time and initial heavy metal ions concentration) on sorption were studied.

The software package Design-Expert 7.0.0 was used to analyse the experimental data and to plot graphs. The correlation of removal of heavy metal ions and four experimental factors were got with coded variables below:

$$\begin{aligned} \text{adsorption capacity Cd}^{2+} = & 12.46 - 0.035A - 7.77B + 0.34C \\ & + 4.51D + 0.78AB - 0.034AC \\ & - 0.54AD + 0.16BC - 1.57BD + 0.13CD \\ & - 0.14A^2 + 3.85B^2 + 0.51C^2 + 0.12D^2 \end{aligned} \quad (3)$$

$$\begin{aligned} \text{adsorption capacity Hg}^{2+} = & 12.43 + 0.83A - 4.75B + 0.11C \\ & + 4.11D - 0.35AB + 0.51AC \\ & + 0.26AD + 0.58BC - 1.4BD - 0.39CD \\ & - 0.51A^2 + 1.57B^2 - 0.086C^2 - 0.074D^2 \end{aligned} \quad (4)$$



**Fig. 5** Effect of factors on sorption for a, d  $\text{Cd}^{2+}$  b, e  $\text{Hg}^{2+}$  c, f  $\text{Cu}^{2+}$

$$\begin{aligned} \text{adsorption capacity Cu}^{2+} = & 16.67 - 0.18A - 7.43B - 0.18C \\ & + 4.44D + 0.2AB - 0.58AC \\ & + 0.24AD - 0.2BC - 0.98BD + 0.2CD \\ & - 0.31A^2 + 3.03B^2 - 0.31C^2 - 0.31D^2 \end{aligned} \quad (5)$$

After studied analysis of variance for removal of  $\text{Hg}^{2+}$ ,  $\text{Cu}^{2+}$  and  $\text{Cd}^{2+}$ , it is found that the quadratic model was good for these results. Lack of fit for three ions were low and not significant that indicated data were consented with the quadratic model. The coefficients of determination ( $R^2$ ) for removal of  $\text{Hg}^{2+}$ ,  $\text{Cu}^{2+}$  and  $\text{Cd}^{2+}$  were 0.97, 0.95 and 0.96, respectively.

3.3. Effect of factors, optimised conditions: Fig. 5 shows the effect of factors on sorption that time and pH have the lowest effect, unlike, adsorbent dosage and concentration analyte have the effect.

Finally, experimental conditions optimised for simultaneous adsorption of three heavy metal ions, where pH, removal time, dosage adsorbent and initial concentration equal 5.0, 20.00 min, 20.00 mg and 25.00 ppm, respectively.

3.4. Equilibrium isotherm models: The adsorption isotherm indicates how molecules of adsorbate are partitioned between the adsorbent and liquid phase at equilibrium as a function of adsorbate concentration [26, 27]. In this Letter, the equilibrium data obtained for the adsorptions of  $\text{Hg}^{2+}$ ,  $\text{Cu}^{2+}$  and  $\text{Cd}^{2+}$  ions onto modified SBA-15 were analysed by considering the Langmuir, Freundlich and Tempkin isotherm model.  $R^2$  values show that the Langmuir isotherm was good for adsorption of heavy metal ions.

**Table 3** Parameters of adsorption isotherms

	Langmuir		Freundlich		Tempkin	
	$Q_m, K_L$	$R^2$	$nK_F$	$R^2$	$B, K_T$	$R^2$
Cu	21.739	1.1219	4.52	1.059	1.25	1.2
	0.999	—	0.832	—	0.941	—
Cd	18.181	1.1702	5.12	1.005	6.151	42.03
	0.999	—	0.834	—	0.847	—
Hg	23.809	1.0243	4.11	1.079	9.188	20.53
	0.921	—	0.998	—	0.899	—

**Table 4** Parameters of the kinetic isotherm

	Pseudo-first order			Pseudo-second order		
	$Q$	$K_1$	$R^2$	$Q_e$	$K_2$	$R^2$
Cu	9.77	0.002	0.99	12.048	0.0246	0.998
Cd	12.317	0.001	0.998	13.333	0.0625	0.998
Hg	15.75	0.002	0.997	18.867	0.0173	0.999

**Table 5** Report for adsorption of  $\text{Hg}^{2+}$ ,  $\text{Cu}^{2+}$  and  $\text{Cd}^{2+}$  in real samples

Sample	Initial concentration, ppm			Added, ppm	Residual concentration, ppm		
	$\text{Hg}^{2+}$	$\text{Cu}^{2+}$	$\text{Cd}^{2+}$		$\text{Hg}^{2+}$	$\text{Cu}^{2+}$	$\text{Cd}^{2+}$
river water	0.00	0.00	0.00	25.00	3.63	2.45	1.74
shaft water	0.00	0.00	0.00	25.00	4.04	2.01	0.98
tap water	0.00	0.11	0.00	25.00	1.87	4.06	4.65

**Table 6** Comparison of adsorption capacities  $Q_{\max}$  (mg g<sup>-1</sup>) of various adsorbents for heavy metal ions

Adsorbent	$Q_{\max}$ , mg g <sup>-1</sup>			Contact time, min	References
	Hg	Cd	Cu		
montmorillonite-kaolinite/TiO <sub>2</sub> composite (simultaneous adsorption)	—	13.8	42.9	60	[28]
activated carbon cloth (simultaneous adsorption)	—	21.86	—	—	[29]
silica-3-chloropropyltriethoxysilane	48.14	—	—	—	[30]
silica-3-(2,2'-dipyridylamine) (DPA)	—	—	82.60	14	[31]
magnetic graphene oxide (GO) (simultaneous adsorption)	—	5.34	18.37	—	[32]
amino and thiolate-functionalised multiwall carbon nanotubes	85	—	—	60	[33]
GO-polydopamine	—	33.3	24.4	—	[34]
SBA-15/DZ	25.04	30.3	31.79	20	this work

The Langmuir (6) was explained by

$$q_e = \frac{C_e K_L Q_m}{1 + K_L C_e} \quad (6)$$

$q_e$  is the equilibrium adsorption capacity (mg/g);  $C_e$  is the equilibrium concentration (mg/l);  $Q_m$  is the maximum adsorption capacity (mg/g);  $K_L$  is the Langmuir constant (l/mg); and  $Q_m$  and  $K_L$  obtained from the slope and intercept of the  $C_e/q_e$  versus  $C_e$  curve, respectively (Table 3).

**3.5. Adsorption kinetics:** A kinetic study of adsorption is necessary as it provides information about the adsorption mechanism, which is crucial for the practicality of the process. In this Letter, two different kinetic models were applied in order to establish which of them shows the best fit with experimentally obtained data.

$R^2$  values in Table 4 indicate that the adsorption process obeyed pseudo-second-order kinetic isotherm in the equation below:

$$\frac{t}{Q_t} = \frac{1}{K_2 Q_e^2} + \frac{t}{Q_e} \quad (7)$$

$t$  is the time (min);  $K_2$  is the pseudo-second-order constant (g/mg min);  $Q_e$  is the equilibrium adsorption capacity (mg/g); and  $Q_t$  is the adsorption capacity at any time (mg/g).

**3.6. Adsorption of real samples and recovered:** According to the results in Table 5, SBA/[M(DZ)<sub>2</sub>] was good for treatment of real samples such as river water, shaft water and tap water. SBA/[M(DZ)<sub>2</sub>] were of sorption ability by high adsorption capacity.

The adsorbent was recovered by 0.10 M HCl and 0.10 M ethylenediaminetetraacetate (EDTA) solution, and the per cent removal remained around 80.7% in the fourth cycle.

The comparison between the performances of SBA-15/DZ adsorbent for simultaneous adsorption of Hg<sup>2+</sup>, Cu<sup>2+</sup> and Cd<sup>2+</sup> ions with various adsorbents is shown in Table 6. As shown, the sum of adsorption capacity for SBA-15/DZ for simultaneous adsorption (87.13 mg g<sup>-1</sup>) is higher than other simultaneous adsorbents.

**4. Conclusions:** In this Letter, the mesoporous silica SBA-15/DZ was synthesised by the hydrothermal method to remove heavy metal ions. Interaction [M(DZ)<sub>2</sub>] complex with mesoporous silica SBA-15 was suitable for simultaneous removal of Hg<sup>2+</sup>, Cu<sup>2+</sup> and Cd<sup>2+</sup> heavy metal ions in aqueous solution and environmental samples, because SBA/[M(DZ)<sub>2</sub>] have more activated adsorption sites than SBA-15 and DZ singly. This sorbent has high adsorption capacity and selectivity, rapid sorption/desorption kinetic and recovered easily by EDTA and HCl solution. The equilibrium and kinetic studies were investigated for the adsorption process. The isotherm models such as Langmuir, Freundlich and Tempkin

were evaluated and the equilibrium data were best described by the Langmuir model. The process kinetics can be successfully fitted to the pseudo-second-order kinetic model.

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